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Microparasitic disease dynamics in benthic suspension feeders: infective dose, non-focal hosts, and particle diffusion

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9 Abstract

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Benthic suspension-feeders can accumulate substantial numbers of microparasitic pathogens by contacting or filtering particles 10 while feeding, thus making them highly vulnerable to infectious diseases. The study of disease dynamics in these marine organisms 11 requires an innovative approach to modeling. To do so, we developed a single-population deterministic compartmental model 12 adapted from the mathematical theory of epidemics. The model is a continuous-time model, unstructured in spatial or age terms, 13 and configured to simulate the dynamics of diverse dose (body burden)-dependent infectious disease transmission processes in 14 suspension feeders caused by susceptible individuals contacting or absorbing (filtering) infectious waterborne pathogens. Different 15 scenarios were simulated to explore the effect of recruitment, filtration rate, particle loss, diffusion-like processes in the water 16 column and non-focal hosts (i.e. non-susceptible in terms of disease) on disease incidence. An increase in recruitment (i.e. new 17 disease free susceptibles) can reduce the prevalence of infection due to the dilution effect of adding more susceptibles, but the 18 disease can spread faster for the same reason. Lower infective particle accumulation rates or increasing particle loss rates in the 19 environment reduce the prevalence of infection. This effect is trivial when the water is saturated with infective particles released 20 by infected and/or dead animals. Diffusion of particles from the local pool available to suspension feeders to the adjacent remote 21 pool, prompted by a large remote volume and high particle exchange, limits epizootic development. Similarly, the likelihood of an 22 epizootic can be constrained in a large susceptible population when competition for pathogens, more 'active' in active filter feeders 23 than in passive suspension feeders, reduces the per capita infective particle accumulation rate. In passive suspension feeders, 24 decreasing the area of the feeding surface has the same effect in constraining disease development. The effect of competition for 25 infective particles in essence diluting the infective particle concentration in the water column is magnified when the susceptible 26 population is part of a community with non-focal filter feeders, and is particularly effective in limiting disease development in 27 high infective dose systems. At the same time, this active foraging strategy makes filter feeders more vulnerable to epizootics. 28 The model is a suitable framework for studying the disease dynamics and determinants of disease outbreaks in benthic suspension 29

30 feeders.

31 Keywords:

32 disease model, epizootic, transmission, suspension feeders, filter feeders, infective-dose, overfiltration, non-focal host

1. Introduction 33

34 Benthic suspension-feeders are among the optimal foragers in the marine context because the energetic cost for 35 capturing prev is almost nil in passive sessile animals and very low in active filter feeders (Riisgård et al., 1993; 36 Riisgård and Larsen, 1995). Suspension feeders play a major role in the structure and function of marine ecosys-37 tems (Dame, 1993; Newell, 2004) by transferring energy from the pelagic zone to the benthos (Newell, 2004; Porter 38 et al., 2004). This feeding mode as a foraging strategy has a downside in terms of the transmission and spread of 39 waterborne diseases. In addition to food particles, suspension-feeders can also accumulate a substantial number of 40 infectious pathogens such as bacteria, fungi, protozoans, and viruses from the water. For instance, the scleractinian 41 coral *Madracis mirabilis* can accumulate 1×10^7 bacterial cells cm⁻² h⁻¹ and clearance rates in bivalves can be much 42 higher (e.g., 8 L h⁻¹ in oysters, Powell et al. (1992)). This potential to accumulate particles, in turn can catalyze major 43 epizootics and massive mortalities causing substantial changes to ecosystem structure and serious losses in shellfish-44 eries and aquaculture (Lafferty et al., 2015). Among the best characterized, geographically wide-spread, and virulent 45 infectious diseases in suspension feeders are MSX and Dermo in oysters (Villalba et al., 2004), white plague disease 46 and black band disease (BBD) in corals (Sokolow et al., 2009; Zvuloni et al., 2009), and brown ring disease BRD, 47 and Perkinsosis in clams (Paillard, 2004; Dang et al., 2010). 48 Proliferation-based disease models have been considered sufficient to describe disease impact in benthic suspen-49 sion feeder populations characterized by rapid non-point-source transmission, such as bivalves (Calvo et al., 2001; 50 Powell et al., 2011, 2012b; Powell and Hofmann, 2015). Host proliferation models alone or together with hydro-51 dynamic models can explore the effect of environmental factors, such as temperature and salinity, on the process 52 of pathogen proliferation and, consequently, on the infection intensity. These models assume high prevalence and 53 simulate epizootic development and host morbidity and mortality based upon population infection intensity as the 54 pathogen proliferates within the host. For these diseases the dynamics of transmission are typically poorly described 55 (Ford, 1992; Ford and Smolowitz, 2007; Gray et al., 2009) and cursorily integrated into disease models (Powell et al., 56 1996, 1999). Only a few studies have adapted the epidemic SIR models (Kermack and McKendrick, 1927; Anderson 57 and May, 1981) to benthic marine organisms, despite the fact that the structure of these models as commonly used for 58 terrestrial diseases is equally applicable and adaptable to describe the epizootiology in benthic animals (McCallum 59 et al., 2004). For this purpose, however, some marine system-specific features such as the buoyancy and the high 60 potential of dispersion and dilution of waterborne pathogens (Strathmann, 1990; McCallum et al., 2003) and certain 61 unique feeding modes such as suspension feeding may need to be incorporated. Kuris and Lafferty (1992) developed a 62 model incorporating both recruitment of the parasite and the host to compare the effect of various management strate-63 gies on a hypothetical crustacean parasitized by a parasitic castrator. In this non-point-source host-pathogen system, 64 the number of nearby infected animals is relatively unimportant in comparison with the number of infective pathogens 65 in the water. McCallum et al. (2005) modeled the dynamics of withering syndrome in abalones by incorporating the 66 free-living pathogen stage and disease transmission through contact between this stage and the host. More recently, 67 Sokolow et al. (2009) and Yakob and Mumby (2011) formulated the dynamics of disease in corals describing the 68 transmission of disease by contact between the host and free-living pathogens. (Bidegain et al., 2016)formulated sim-69 ple compartmental models to yield the basic reproduction number R_0 for a variety of marine host-pathogen systems 70 to explore the relative importance of the host and pathogen traits that determine transmission. 71 Most compartmental disease models (e.g., SI, SIR, SEIR, etc.) follow the classic mass action approach where 72 disease transmission is a function of the number of contacts between susceptible individuals and infective particles 73 in the water (Regoes et al., 2002; Sokolow et al., 2009; Yakob and Mumby, 2011). However, it may be important to 74 include the well-known dose-effect in disease transmission for some suspension feeders such as oysters (Bushek et al., 75 1997; Ford et al., 1999; Powell et al., 1999); effective dose is a characteristic with potentially interesting effects on 76 disease transmission that seems crucial to take into account. Active filter feeders can be sufficiently dense to compete 77 for food (e.g., Fréchette et al., 1992; Wilson–Ormond et al., 1997; Widdows et al., 2002) particularly under conditions 78 of slow flow and limited vertical advection. Similarly, they may 'compete' for pathogens, presumably reducing the 79 concentration of infective particles sufficiently to limit body burden below the infective dose and, in turn, limiting 80

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81 epizootic development (i.e. the overfiltration scenario - Bidegain et al. in press). Passive filter feeders may show this

82 mechanism less obviously since they more importantly depend on the ambient flow and the body surface exposed to that flow, rather than on an active pumping of water from the environment through a filter (Sebens et al., 1996, 1998). 83

When host diversity increases, the disease risk can decrease in what is known as the dilution effect. This wellstudied effect in Lyme disease (Ostfeld and Keesing, 2000) appears to be a more general phenomenon that relies on the idea that a certain community with high species diversity will contain a proportion of incompetent hosts, in terms of disease, that compete for particles and deflect infectious pathogens away from the susceptible hosts, thereby reducing

⁸⁸ infection prevalence and disease risk. This marine suspension feeder community could be composed, for instance,

⁸⁹ by the scleractinian coral *Madracis mirabilis* and the colonial ascidian *Trididemnum solidun*, since data show that

⁹⁰ both organisms are effective bacterial suspension feeders (Bak et al., 1998). Other examples may by a reef composed

⁹¹ by oysters and mussels, since mussels that attach to oyster reefs and shells can double the reef's filtration capacity

⁹² (Gedan et al., 2014), or by oysters and ascidian tunicates (T. Ben-Horin, unpublished data).

Another interesting characteristic of some diseases of suspension feeders is the importance of dead infected ani-93 mals as a source of infective particles and, hence, the role of mortality and subsequent scavenging and tissue decay in 94 disease transmission. For instance, the pathogen body burden in dead oysters infected by Dermo disease is commonly 95 higher and the potential release rate upon death is relatively much faster than that of infected live animals (Bushek 96 et al., 2002). Dead corals infected by the black band disease may also release pathogens through breakdown of decay-97 ing tissue (Richardson, 2004). Once released, diffusion processes in the water column may have an effect in pathogen 98 dynamics and consequently, in disease transmission. The population turnover rate (i.e. high mortality and recruitment 99 rates) thus is another mechanism of controlling epizootics in suspension feeders. Yakob and Mumby (2011) found that 100 allowing for a more dynamic population turnover in an epizoological model of coral disease not only gives a superior 101 fit to empirical data, but also suggests that emerging coral assemblages could be far less prone to epizootics. The role 102 of predation in the context of SIR models has been considered (Su and Hui, 2011; Wang et al., 2011), but has rarely 103 been applied in the marine context (Liao et al., 2008). In filter feeders such as oysters the important contribution of 104 dead animals releasing infective particles may counterweigh the effect of the population turnover rate as a restraining 105 influence on epizootics (Bidegain et al., unpublished). 106

A theoretical transmission-based model or a host-population model that incorporates all of these features and de-107 scribes the dynamics of the host in all possible stages (i.e. susceptible, live infected, dead infected) and the waterborne 108 pathogens does not exist for active or passive suspension feeders. In this paper, we develop an SI model capable of 109 reproducing these processes. For this purpose, we adapt the Kermack and McKendrick (1927) epidemiological the-110 ory and the microparasitic epidemic model of Anderson and May (1981). The model reproduces the dynamics of 111 both susceptible and infective stages of the host, whether alive or dead, and the waterborne infective pathogens. Our 112 study system includes dose-dependent (i.e. body burden dependent) disease transmission wherein suspension feeders 113 contact or absorb (filtering) waterborne infective pathogens released by live or dead infected animals. The model con-114 templates the effect on disease transmission of non-point sources of pathogens and diffusion processes in the water 115 column. Moreover, the model permits the study of potential mechanisms by which suspension feeders might dilute 116 the disease risk, such as competitive interactions among and between species, and the effect of recruitment of new 117 susceptible individuals. Finally, the model yields the formulation of the basic reproduction number R_0 for both active 118 and passive filter feeders, to comprehensively describe the relative importance of some of these processes driving 119 epizootics. 120

121 2. The model

122 2.1. Mathematical theory and model structure

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The model is a one-population deterministic compartmental model continuous in time, unstructured in spatial or age terms (Cuddington and Beisner, 2005), and configured to simulate the dynamics of a diversity of infectious disease transmission processes in suspension feeders caused by susceptible individuals contacting or absorbing (filtering) infectious waterborne pathogens such as bacteria, fungi, protozoans, and viruses from the water column.

As a compartmental model, which is the most frequently used class of model in epidemiology (Diekmann et al., 2013), individuals and pathogens can take on a finite number of discrete states. Each state is representative of a subpopulation of individuals or pathogens at a given time (Table 1) which, in turn, together with the corresponding parameters (Table 2) satisfy a system of nonlinear ordinary differential equations (ODEs) describing the dynamics of the host-pathogen association. The incorporation of a dead infected subpopulation stems form the fact that some suspension feeders such as oysters or corals may also transmit the disease by releasing pathogens upon death (Bushek
 et al., 2002; Richardson, 2004). A variant of the model incorporates another variable representing an alternate suspen sion feeder host population which is incompetent in terms of disease (*H*), that is, the alternate host does not develop
 the disease and pathogens are inactivated inside the animals acting as an important mechanism of controlling the
 concentration of infective particles in the water column and as a sink of pathogens.

The population of waterborne pathogens is divided into three classes depending upon their location in the system 138 based on the division of the environment into two volumes of water: The first is defined as the 'local pool' of pathogens 139 P. This pool is the pathogen concentration in the 'local volume' defined as V_l , adjacent to the bottom, within which 140 pathogens are released by infected and dead infected individuals and remain free floating. Pathogens in this volume 141 are susceptible to contact with or being filtered by hosts, and can lose their infective properties after some time or 142 otherwise be lost (i.e advection, diffusion, predation). For example, for oysters, the height of the local volume or the 143 volume available for filtration is considered to be around 10 cm (Wilson-Ormond et al., 1997). This is a typical height 144 for an oyster clump (Powell et al., 1987, 2012a), but this height would exaggerate the effect of an infaunal filter feeder 145 (Ertman and Jumars, 1988; Monismith et al., 1990; Widdows et al., 1998). The second state variable for pathogens 146 is the remote pool U. This second pool is the concentration of pathogens in a second volume contiguous to the local 147 volume, defined as the remote volume V_r , where a remote pool of particles can accumulate without direct interaction 148 with hosts. The size of the remote volume would depend on how deep the water column is above the near-bottom layer 149 directly affected by the filtering benthic population. Diffusional exchange of particles between these two volumes is 150 assumed. The third subpopulation of pathogens is that accumulated in the susceptible population through contact or 151 filtration (F). 152

The infection state in the model is presence/absence, that is, the animals are infected or are not, and if they are infected, they have the 'average' parasite load. We assume this because microparasites are usually (but not always) unicellular microorganisms, such as viruses, bacteria and protozoans that can multiply rapidly within a host (McCallum et al., 2004) relative to a one year time step. The transfer of individuals from absence to presence of disease is by a dose-dependent or body burden dependent transmission process detailed in section 2.4.1.

158 2.2. Miscellany

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Variables related to the host population are defined with respect to the concentration of individuals in a given surface area of the bottom, whereas variables related to the pathogens are defined in number of particles in a given water volume or the number of particles *in vivo*. Note that for simplicity in rendering equations, the reciprocals of the local and remote volumes V_l and V_r , *sl* and *sr* respectively, are often used.

Variable	Definition	Unit		
S	Susceptible hosts in the population	Number of individuals		
Ι	Infected individuals in the population	Number of individuals		
DI	Dead infected individuals in the population	Number of individuals		
DS	Dead susceptible individuals in the population	Number of individuals		
Р	Free-living pathogens in the environment, local pool	Number of particles		
F	Total number of pathogens absorbed or filtered by the population	Number of particles		
U	Pathogens from a remote volume, remote pool	Number of particles		
Н	Alternate non-competent reservoir host	Number of individuals		

Table 1: Variables of the model. Note that the model has an implicit surface area or volume for the host subpopulations and pathogens, respectively. *F* is the exception and the corresponding unit is internal particles in the susceptible population.

The mathematical theory of this model contemplates that disease transmission requires the pathogen to exist outside the host and remain infectious for sometime finite period. Thus, the susceptible and infected individuals are

not necessarily in contact or even components of the same local population, since the free-living pathogen stage can 166 drift in the fluid and contact or be filtered by susceptible animals located near or far away from the particle source. 167 Natural mortality of the host and mortality due to disease are also integrated into the formulation of the model. Thus, 168 the infected individuals can die due to both natural causes and disease. The model also assumes an open population 169 since it contemplates demographic turnover; that is, the recruitment process is integrated in the dynamics of the 170 host. The model does not incorporate the process of recovery from disease because only a few examples of marine 171 wild populations recovering from disease are known in marine systems (Gilmour et al., 2013; Paillard et al., 2014; 172 Vega Thurber et al., 2014). Consequently, individuals never recover from the disease and infected individuals remain 173

¹⁷⁴ infected until they die.

175 2.3. Model scheme

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The flow diagram of the model (Fig. 1) shows the most important processes in the disease transmission process in 177 suspension feeders. Disease is transmitted to the susceptible population S thorough filtration (active suspension feed-178 ers) or contact (passive suspension feeders) at a rate β by a dose-dependent transmission. Infected animals I die due 179 to both natural mortality m_S and disease mortality m_I , while susceptible individuals S only die due to natural causes. 180 The alternate incompetent (i.e. not susceptible to the disease) suspension feeder host H competing for waterborne 181 pathogens with the susceptible host is also located in the bottom. This alternate host only dies by natural mortality 182 assuming that it is resistant to the disease and does not release particles to the water; that is, they are assumed to be 183 inactivated by the immune system or by diapedesis. An exchange between the local near-bottom pathogen pool P 184 and the remote pool U occurs at a certain exchange rate γ by a diffusion-like process proportional to the difference in 185

concentration between the two pools.



Figure 1: Model flow diagram. The model variables are represented by capital letters: susceptible animals (S), infected animals (I), dead susceptible (DS) and dead infected animals (DI), waterborne pathogens (P), internal pool of pathogens (F), remote pool of pathogens (U) and alternate host population (H). Dashed arrows represent the main processes in the model. The parameters involved in these transmission are described in Table 2.

187 2.4. Model equations

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The equations that follow represent the disease transmission process and the model variables. The host and pathogen states or subpopulations satisfy a system of ODEs describing the dynamic of the host-pathogen association. The numerical model for this ODE system is programmed in Matlab 8.1. The set of coupled differential equations is solved with a 4^{th} -order predictor corrector scheme using the Adams-Bashforth predictor and the Adams-Moulton corrector.

¹⁹⁴ 2.4.1. Infective dose and disease transmission

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Disease transmission in suspension feeders, at least in filter feeders, most likely occurs via an infective dose (Bushek et al., 1997; Ford et al., 1999; Powell et al., 1999) rather than by unique contact between a single pathogen or fomite and host. Consequently, the transmission process in the model is based on the fact that susceptible animals require some minimum level of body burden of infectious particles, the infective dose, to become infectious. Furthermore, the model assumes that as hosts absorb or filter particles, some susceptible individuals will have a relatively large body burden whereas most will have a smaller body burden. The distribution of the number of susceptible animals (S) with each level of body burden (b) (Fig. 2) is assumed to have the form

$$S(b) = S_0 e^{-\rho b},$$
 (1)

which has a simpler expression but similar behavior to the negative binomial distribution, very common in benthic infected hosts (Anderson and Gordon, 1982; Ford et al., 1999)), in having a long tail describing the declining frequency of individuals of increasing infection in the population distribution. S_0 is the total number of susceptible individuals. For body burdens up to several hundred, ρ is a well approximated by \hat{S}/F , where \hat{S} is the number of susceptible animals and F is the number of infectious particles housed in the susceptible subpopulation (Eq. (2)). Thus F/Srepresents the average concentration of infectious particles in susceptible individuals.

$$\frac{dF}{dt} = -(\beta_F + m_S + a + c_S)F + f_S A_S S P.$$
⁽²⁾

Infectious particles are filtered out (active filter feeders) by or come into contact (passive filter feeders) with 209 susceptible individuals at the filtration or contact rate f_S , being the contact rate proportional to the feeding surface A_S 210 for passive filter feeders (see section 2.4.2 for a more detailed description of particle uptake). Infectious particles can 211 be removed from the susceptible population by four processes: (i) the reduction of the internal pool of particles in the 212 susceptible population due to removal of susceptible individuals that become infected t_F (see details in Eq. (6)), (ii) 213 the background mortality m_S as it removes individuals with incorporated infected particles, (iii) the inactivation inside 214 susceptible animals (a), controlled by some component of the immune response such as phagocytosis, inactivation by 215 oxygen radicles, binding by lectins, etc. (Renwrantz, 1986; Villalba et al., 2004), and (iv) the release of particles from 216 susceptibles through, for instance, faeces. Whether a susceptible becomes infected is a balance between a, the release 217 rate of particles c_s , the filtration of particles by susceptibles $f_s S P$, and the concentration of infective particles that 218 needs to be reached to generate the infective dose (b_{min} , see Fig. 2). 219

The distribution of infective particle body burden in susceptible hosts in Fig. 2 has a very long tail so that a few uninfected animals have a high body burden. The portion of the susceptible pool eligible to transition to the infected pool is circumscribed by b_{max} , the maximum body burden for any animal considered to be uninfected and b_{min} , the minimum body burden for animals eligible to transition to the infected pool. At any time, there will be some number of susceptible animals (*S*) with a total absorbed pool of infectious particles (*F*). These values are used to determine how many susceptible animals with a body burden between b_{min} and b_{max} should become infected.

²²⁶ The total number of susceptible animals at any time is

$$S = \int_0^{b_{max}} \hat{S}_0 \, e^{-\rho \, b} \, db = \frac{\hat{S}_0}{\rho} (1 - e^{-b_{max}}) \tag{3}$$

where \hat{S}_0 is the number of susceptibles per body burden. The total body burden for the population is

$$F = \int_0^{b_{max}} \hat{S}_0 \ b \ e^{-\rho \ b} \ db = \frac{\hat{S}_0}{\rho^2} \left(1 - (1 + \rho \ b_{max}) \ e^{-\rho \ b_{max}}.$$
(4)

²²⁷ F and \hat{S} are both known at any step in the model. Consequently, the values of S_0 and ρ can be calculated easily. ²²⁸ The total number of susceptible individuals moved to the infected population is



Figure 2: A theoretical distribution of the population with each level of infective particle body burden (b). The minimum body burden at which susceptible animals are considered to be infected and thus transfered to the infected population is b_{min}. The maximum body burden for any susceptible animal considered to be uninfected is b_{max} .

$$t_{S} = \int_{b_{min}}^{b_{max}} \hat{S}_{0} e^{-\rho b} db = \frac{\hat{S}_{0}}{\rho} (e^{-\rho b_{min}} - e^{-\rho b_{max}}).$$
(5)

The total number of infectious particles that should be removed from the susceptible internal pool as a consequence is 229

$$t_F = \int_{b_{min}}^{b_{max}} b\,\hat{S}_0\,e^{-\rho\,b}\,db = \frac{\hat{S}_0}{\rho^2}((1+\rho\,b_{min})\,e^{-\rho\,b_{min}}) - (1+\rho\,b_{max})\,e^{-\rho\,b_{max}}). \tag{6}$$

Note that since the heavily infected susceptibles are the ones that are moved to the infected pool, a large number 230 of internal infectious particles will be removed from the internal pool of the susceptibles. Based on these calculations, 231 a fraction t_S/S of the susceptibles need to be moved to the infected pool and t_F/F of the internal infectious particles 232 need to be removed from the susceptible internal pool. These changes are assumed to occur at a certain rate α [1/day] 233 depending on the host-pathogen system. For instance, for a change that occurs at a rate of 10% per day, α is 2.3. The 234 terms in the governing equations representing disease transmission thus have the form 235

$$\beta_S = \alpha \, \frac{t_S}{S} \tag{7}$$

where α is a specific rate to move a fraction of the individuals from the susceptible population S to the infected 236 population I. Similarly, the rate at which the internal particles are removed from the internal pool F is 237

$$\beta_F = \alpha \, \frac{t_F}{F}.\tag{8}$$

Note that the standard mass action model with infective particles with an instantaneous dose response, typically 238 represented as βPS (Bidegain et al., in press), is a specific case of the dose-response model presented here. For $\rho = 0$, 239 Eq. (5) becomes 240

$$t_S = \hat{S}_0 \left(b_{max} - b_{min} \right), \tag{9}$$

and consequently, $S = \hat{S}_0 b_{max}$ and $\beta_S = \alpha (1 - \frac{b_{min}}{b_{max}})$. The standard mass action model does not explicitly specify particle accumulation or body burden, so that $b_{min} = 0$; thus, Eq. 7 simplifies for the case of an instantaneous dose 241

242 243 response transmission $\beta_S = \alpha$.

244 2.4.2. Uptake of particles by filtration or contact: active and passive suspension feeders

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The model is formulated to permit the study of both active (e.g., bivalves) and passive (e.g. corals) suspension 246 feeders. The main feeding difference between these organisms is that active filter feeders pump water actively through 247 a sieve while passive suspension feeders are infected by passively contacting particles transported by water currents. 248 Thus, disease transmission in filter feeders depends on the specific filtration rate (Powell et al., 1999) whereas, for 249 passive suspension feeders, the transmission of disease is a function of the exposed surface area of the individual or 250 colony devoted to food collection and the water flow speed [Sebens et al. (1996) and Sebens et al. (1998) demonstrated 251 that food capture is higher in corals with high polyp and colony exposed surface area in relation to their biomass, and 252 is limited at low flow conditions]. The model incorporates the term f to represent these two foraging strategies. In 253 both cases, we consider that the capture rate of food and infective particles is similar. Consequently, for active filter 254 feeders f represents the rate at which particles are filtered by one individual per unit of time, whereas for passive 255 suspension feeders f represents the rate at which the particles contact the exposed surface area of an individual per 256 unit of time (see Table 2). 257

259 Active filter feeders

Particle removal from the water due to active filtration by susceptible individuals (see Eq. (14)) has the form $f_S S P$, where f_S is the filtration rate (volume filtered) per susceptible individual, S is the number of benthic susceptible individuals, and P is the concentration. Similarly, the term to represent the particle removal due to active filtration of infected individuals (I) and non-focal or alternate incompetent hosts in terms of disease (H) would be $f_I IP$ and $f_H HP$, respectively. The incorporation of specific filtration rates for susceptible and infected individuals permits simulation of disease dynamics for cases where infected or severely diseased individuals show a reduction in clearance rate. Similarly, the non-focal host has a specific filtration rate.

268 269 Passive suspension feeders

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For passive suspension feeders, the particle removal from the water by passive contact (see Eq. (15)) has the form 271 $f S A_S P$, where f S is the adrift particle contact rate per susceptible individual and A_S is the exposed surface area 272 per individual. Similarly, the term to represent the particle removal due to passive contact of infected individuals (I) 273 and non-focal or alternate incompetent hosts (H) with particles would be $f_I I A_I P$ and $f_H H A_s P$, respectively. The 274 model also specifies specific contact rates for susceptible and infected individuals, and for non-focal hosts. However, 275 considering that (i) the water movement is the primary mechanism bringing zooplankton into contact with passive 276 suspension feeders and that exposed surface is crucial (Sebens et al., 1998), and (ii) the model already incorporates a 277 term for individual exposed surface area, the term f may be a function of the water flow speed rather than a disease-278 affected characteristic and, thus, may likely be considered the same for different subpopulations of the same host. 279

280 2.4.3. Susceptible Individuals

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²⁸² *S* is the number of susceptible animals in a given surface area of the bottom. Susceptible individuals are lost by ²⁸³ two processes: infection and natural mortality. The disease transmission rate is controlled by β_S which is a function ²⁸⁴ of body burden, allowing for a dose-response to infection, and it is an estimate of the rate at which individuals become ²⁸⁵ infected (see details in section 2.4.1), whereas the natural mortality rate is m_S (Eq. (10)).

The model allows recruitment of new susceptible individuals when the population density (S+I) falls below the carrying capacity *K*. Susceptible and infected animals can produce recruits at different rates n_S and n_I , respectively allowing for reduced recruitment associated with infected individuals (Yakob and Mumby, 2011). Finally, the model permits an external source of recruits (*SRC*_S).

$$\frac{dS}{dt} = -(\beta_S + m_S)S + \left(1 - \frac{S+I}{K}\right)(n_S S + n_I I + SRC_S)$$
(10)

where, if $1 - \frac{S+I}{K} > 0$, then the carrying capacity is not reached and new individuals can be recruited to the susceptible population as a function of the recruitment rates n_S , n_I , and the potential external source of recruits Src_S .

292 2.4.4. Infected Individuals

Infected individuals are transferred from the susceptible subpopulation to the infected subpopulation at the rate β_S (Eq. (11)). Infected individuals die at a rate controlled by natural mortality m_S and disease mortality m_I .

$$\frac{dI}{dt} = \beta_S S - (m_S + m_I)I \tag{11}$$

294 2.4.5. Dead Susceptible Individuals

²⁹⁶ Susceptible individuals die at a rate m_S which is a background mortality not related to disease processes (Eqs. ²⁹⁷ (10) and (12)). The subpopulation of dead individuals is reduced at a rate *d*, which represents bacterial decomposition ²⁹⁸ or scavenging processes. Although true scavengers do not exist in the marine world, many predators scavenge adven-²⁹⁹ titiously (Hoese, 1962; Veale et al., 2000; Morello et al., 2005). These dead susceptible individuals are a diagnostic ³⁰⁰ and do not affect the behavior of the model.

$$\frac{d\,DS}{d\,t} = m_S\,S - d\,DS\tag{12}$$

301 2.4.6. Dead Infected Individuals

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Infected individuals die from disease at a mortality rate m_I and suffer background mortality at rate m_S (Eqs. (11) and (13)). Similarly to dead susceptible animals, these dead infected individuals decay or are scavenged at a rate d.

$$\frac{dDI}{dt} = (m_S + m_I)I - dDI$$
(13)

305 2.4.7. Infectious Particles in the local volume

The benthic community comprises susceptible hosts, infected hosts, and alternate hosts, filtering out or contacting particles. The model analyses two situations in terms of the type of benthic system contributing to the removal of particles from the system.

³¹⁰ 2.4.7.1. Susceptible and infected animals removing and releasing particles.

P is the number of infectious particles in the volume of water immediately accesible to the suspension feeder 312 population (Eqs. (14) and (15)). Infectious particles are mainly added to the water by release from infected individuals 313 at a rate c_I (e.g. by faeces) and from dead infected individuals at a rate c_{DI} . Susceptibles and dead susceptibles can 314 release a small amount of particles through faeces at a rate c_s and upon death through decomposition and scavenging 315 processes at a rate c_{DS} . If dead infected animals release their entire body burden of particles, then $c_{DI} = d$, or they 316 may release a much smaller number depending on the characteristics of the decay or scavenging process removing 317 dead animals, which may inactivate a substantial proportion of or all of the infective particles before they can be 318 released into the water column. Infectious particles are removed from the local volume by one of two processes: (i) 319 inactivation at a rate r, by dilution, transport downstream, or by reduction of infectiousness by inactivation or death, 320 and (ii) filtration by or contact with susceptible and infected individuals. 321

The local water volume can exchange particles with the vertically adjacent volume if the concentrations are different, through a diffusion-like process controlled by parameter γ (see section 2.2) (Eqs. (14) and (15)). Finally, an unspecified source of infectious particles from another benthic community (*SRC_P*) is also permitted.

325

327

326 Active filter feeders

At this point, the model differentiates between active filter feeders (e.g., bivalves) (Eq. (14)) and passive suspension feeders (e.g., corals) (Equation 15) (see also section 2.4.2) to represent the distinct removal of particles from the water. Thus, the change in the number of infective particles in a given water volume for active filter feeders is

$$\frac{dP}{dt} = b_I c_I I + c_S F + b_{DI} c_{DI} DI - (r + f_S S + f_I I + f_H H) P - \gamma (slP - srU) + SRC_P$$
(14)

where the terms for removal of particles due to filtration have the form $f_S S P$ or $f_I I P$, for susceptible and infected individuals respectively, and f_S and f_I are the filtration rate of particles per susceptible and infected individuals, respectively (see units in Table 2). An alternate active filter feeder host population H can also remove particles from the system at the filtration rate f_H (Eq. (14). The alternate host population is assumed to be non-competent in terms of disease. Consequently, the infected and dead individuals of this population do not release infective particles.

336 337 338

Passive suspension feeders

Alternatively, for passive suspension-feeding hosts, the change in the number of infective particles in a given water volume is

$$\frac{dP}{dt} = b_I c_I I + c_S F + b_{DI} c_{DI} DI - (r + f_S A_S S + f_I A_S I + f_H A_H H) P - \gamma (sl P - sr U) + SRC_P,$$
(15)

where the terms for removal of particles due to passive contact of susceptible and infected individuals with particles have the form $f_S S A_S P$ or $f_I I A_S P$, respectively, where A_H represents the particle collection area. An non-focal passive feeder host population H can also contact particles at a rate f_H and function of the particle collection area (Eq. (15)).

345 2.4.8. Infectious particles in the remote volume

346

³⁴⁷ *U* is the number of infectious particles in the remote pool that cannot be directly affected by or affect the host ³⁴⁸ population (Eq. (16)). The incorporation of the remote pool and the exchange of pathogens between pools facilitates ³⁴⁹ the understanding of non-point sources of pathogens and the diffusion processes of waterborne pathogens. The particle ³⁵⁰ exchange γ between the two water volumes describes the process of diffusion of particles. The other process in this ³⁶¹ equation is controlled by σ , which represents the fact that particles may be removed from the system by loss from the ³⁶² remote pool either through mortality, inactivation, or loss from the system via advection (e.g. tidal exchange). The ³⁶³ model also contemplates the fact that there can be an external source of infective particles (*SRC_U*).

$$\frac{dU}{dt} = \gamma \left(slP - srU \right) - \sigma U + SRC_U.$$
(16)

$_{354}$ 2.5. R_0 estimation

355

 R_0 represents the number of new cases of infection caused by one infected individual in a population of only 356 susceptible individuals. The definition of R_0 in an epidemiological context includes the threshold value of 1, wherein, 357 if $R_0 > 1$, an epidemic can occur and if $R_0 < 1$, an outbreak is not expected (Diekmann et al., 1990; Dietz, 1993). 358 We estimate R_0 using the next-generation matrices (NGM) method for compartmental epidemic models following 359 Diekmann et al. (2010, 2013). Since, R_0 represents the new cases of infection due to one infected animal, likely 360 occurring in a relatively short time scale, we consider the 'close population' variant of the disease dynamic model; 361 that is we do not consider recruitment or non-disease mortality in calculating R_0 . The population remains constant, 362 except as far as it is modified by deaths due to the disease itself. The R_0 model considers the two potential infective 363 states: infected individuals and dead infected individuals. We formulate specific R_0 for active and passive filter feeders 364 including the effect of non-focal hosts. 365

366 2.5.1. R₀

367

The basic reproduction number for a system composed of susceptible suspension feeders and non-focal hosts is:

$$R_{0} = \sqrt[4]{\frac{\beta_{S}}{a} \left(\frac{c_{I} b_{I}}{m_{I}} + \frac{c_{DI} b_{DI}}{d}\right) \frac{f_{S} A_{S} N}{r + f_{S} A_{S} N + f_{H} A_{H} H + \frac{\gamma}{V_{I}} \left(\frac{1}{1 + \frac{\gamma}{\sigma V_{r}}}\right)}}.$$
(17)

This equation permits to formulate R_0 for specific systems regarding the feeding behavior of the host (i.e. active or passive suspension feeder) and the presence of non-focal hosts.

371 3. Results

372

The parameter values in Table 2 were used for simulations in sections 3.1 and 3.2 (Figs. 3 and 5). For the rest of the simulations the varying parameter values are specified in each section. Each simulation has a time step of 0.005 day and runs for 1000 days.

376 3.1. Host dynamics

377

We conducted two simulations to evaluate the performance of the model regarding the dynamics of the host 378 population. The first case (Case 1 in Fig. 3) was run with the conditions for parameters given in Table 2, with a 379 transmission coefficient of $\alpha = 0.2$ and the carrying capacity K is 300 individuals. The second case (Case 2 in Fig. 380 4) was run for a lower transmission coefficient ($\alpha = 0.02$) and a higher carrying capacity (K = 500). In both cases, 381 initially there are 300 susceptible benthic individuals and zero infected individuals. Initial conditions include 12,000 382 infective particles m^{-3} in the near-bottom or local volume V_l (Audemard et al., 2006) and none in the remote volume 383 V_r (i.e. U = 0). These simulations also assume a transfer of particles between the two pools, and that particles in the 384 near bottom are removed by animals through filtration or passive contact, or by decay or dilution. Population turnover 385 is function of natural and disease mortality, and recruitment of new susceptible individuals. 386



Figure 3: Case 1. Time evolution of host subpopulations (a) susceptible S and infected I, (b) dead susceptible and dead infected, (c) prevalence of infection, and (d) mortality over 1000 days. Mortality is due to both disease and natural causes, assuming a natural mortality of 10% per year. Parameter values in Table 2

In Case 1 (Fig. 3), the susceptible subpopulation decreases as individuals get infected by filtering or contacting infective particles, while the infected subpopulation increases as susceptible individuals get infected and are transfered to the infected subpopulation (Fig. 3a). In Case 2 (Fig. 4) shows more clearly the initial effect of the recruitment increasing the number of susceptible animals (Fig. 4a) as the population reaches carrying capacity. Moreover, the lower transmission rate in the second case gives a larger final susceptible population with respect to the infected



Figure 4: Case 2. Time evolution of host subpopulations (a) susceptible S and infected I, (b) dead susceptible and dead infected, (c) prevalence of infection, and (d) mortality over 1000 days. Mortality is due to both disease and natural causes, assuming a natural mortality of 10% per year. Parameter values as in Table 2 with the exception of the transmission rate coefficient $\alpha = 0.02$ and the carrying capacity K = 500

³⁹² population (Fig. 4b). The dead susceptible and dead infected subpopulations are small due to the high removal rate of ³⁹³ dead animals, the number being smaller for the second case because of the lower transmission rate and lower disease ³⁹⁴ mortality (Figs. 3b and 4b). In the first case, the number of dead infected animals increases to a maximum in the ³⁹⁵ second year. As a consequence of rapid infection, almost all dead animals died of infection; an insignificant number ³⁹⁶ of susceptible animals died naturally (Fig. 3b).

Due to the high particle concentration and high transmission rate, in Case 1, the prevalence of infection increases 397 until it reaches a steady and high disease prevalence of almost 90% (Fig. 3c). In Case 2, the prevalence is substantially 398 lower (65%) and the steady state is not yet reached at the end of the simulation (Fig. 4c). The mortality is also higher 399 in Case 1 at the end of the simulation (Fig.3d) compared to Case 2 (Fig. 4d). Disease mortality in both cases greatly 400 exceeds the natural mortality rate of 10% per year; thus mortality in both cases is high due to disease. Case 1, in 401 particular, could be representative of a population heavily impacted by disease both in active filter feeders such as 402 oysters (e.g. Dermo and MSX diseases, (Ford and Tripp, 1996; Ford et al., 2006) and passive suspension feeders such 403 as corals (e.g. white plague disease-Yakob and Mumby (2011)). 404

⁴⁰⁵ Considering the diversity of benthic populations with a wide range of transmission rates and carrying capacities, ⁴⁰⁶ in addition to these two specific cases described in detail (Figs.3 and 4), we explore the prevalence of infection for a ⁴⁰⁷ wider range of transmission coefficients α and carrying capacities *K* (Fig. 5). Initial conditions include 1200 infective ⁴⁰⁸ particles m⁻³ in the near-bottom or local volume V_l (a much lower initial number of particles than in Cases 1 and ⁴⁰⁹ 2) in order to better detect and understand the effect of *K* and α on prevalence of infection in suspension feeders. A ⁴¹⁰ very high number of particles in the water column would lead to similarly high prevalence of infections for almost the ⁴¹¹ whole range of values.

Increasing transmission rates leads to higher prevalence of infection in suspension feeders, especially at low carrying capacities (Fig. 5). Increasing carrying capacity can reduce disease transmission even for relatively high



Figure 5: The relationship of prevalence of infection (%) at day 1000 with the transmission coefficient α and the carrying capacity of filter feeders *K*. Carrying capacity is defined in terms of 76-mm oyster equivalents. X-axis values represent a range of values observed in literature accounts. Other parameter values as in Table 2.

disease transmission rates due to the effect of overfiltration and competition for pathogens, particularly in filter-feeder
 high-density populations resulting in lower per capita exposure to pathogens.

416 3.2. Particle dynamics

417

Particle dynamics corresponding to Case 1 with initial conditions and parameter values given in Table 2 are shown 418 in Fig. 6. The total internal number of particles in the susceptible population increases to a maximum as susceptible 419 individuals accumulate particles without reaching the infective (Fig. 6a). This maximum is followed by a rapid 420 decrease as infective doses are reached in the susceptible population. The decrease in the susceptible population is in 421 turn a consequence of the per capita dose reaching the infective dose. Beyond this point, the total accumulation of 422 particles in the population decreases because the number of susceptible individuals in the population decreases until a 423 steady level corresponding to the steady susceptible population level is reached. Resulting steady state occurs when a 424 balance is reached between disease mortality and recruitment (Fig. 6a). 425 The per capita particle body burden in susceptible individuals increases (Fig. 6b) to a maximum as a consequence 426

⁴²⁶ The per capita particle body burden in susceptible individuals increases (Fig. 6b) to a maximum as a consequence ⁴²⁷ of a combination of both a high concentration of particles in the local volume (Fig. 6c) and a still relatively high ⁴²⁸ density of susceptible individuals. This increase in particle in concentration is due to the release of a large quantity ⁴²⁹ of particles from the increasing subpopulations of infected and dead infected animals (Fig. 6a,b). The dead animals ⁴³⁰ release particles more rapidly than live infected animals ($c_{DI} = 0.5$, $c_I = 0.007$) and thus have a more severe effect on ⁴³¹ the prevalence of infection. The low particle loss rate (r = 0.05) due to decay, diffusion, or advection does not balance the release of particle from infected animals. The particle dynamics in the remote pool U have the same pattern, but a slightly lower concentration (Fig. 6d) resulting from the diffusion-like process between the two pools.



Figure 6: Time evolution of infective particles over 1000 days of simulation: (a) internal particle pool in the susceptible population, (b) internal particles per susceptible individual, (c) concentration of infective particles in the volume available for the benthic population (local pool), (d) concentration of infective particles in the remote volume adjacent to the local volume. Parameter values in Table 2.

434 3.3. Factors affecting prevalence of infection

435

We conducted four simulations to evaluate the effect of varying (i) recruitment rate $(n_{S,I})$, (ii) particle filtration or contact rate $f_{S,I}$, (iii) particle loss rate in the local pool (*r*) and (iv) rates of diffusion of particles by varying the remote pool V_r):local volume (V_l) size ratio, on the prevalence of infection (Fig. 7). Non-varying parameter values are given in Table 2, except that a lower transmission rate β_S was used (i.e. $\alpha = 0.02$).

440 3.3.1. Recruitment effect

441

Three recruitment intensities were simulated to compare the relative effect of recruitment rate on the prevalence of 442 infection (Fig. 7a). The simulations with the population at carrying capacity (K = 300) and a transmission rate of β_S 443 for a transmission coefficient $\alpha = 0.02$ showed that increasing recruitment intensity limits the prevalence of infection 444 beyond 150 days of simulation with a maximum reduction of 20% between the lowest and highest recruitment sce-445 narios. The similar increase in prevalence of infection at the beginning of the simulation is explained by the fact that 446 447 the mortality due to disease is too low initially (Fig. 3d) for the population to drop significantly below the carrying 448 capacity. That is, the effect of recruitment on prevalence of infection is visible and increases with the number of new susceptible recruits, which occurs as susceptibles become infected, die and are removed from the system. Together 449 with a decrease in prevalence of infection due to the addition of new individuals, is an increase in the number of 450 individuals becoming infected (Fig. 8b). The addition of individuals by recruitment can also result in a more dense 451

⁴⁵² population filtering out and competing for pathogens, thus reducing the concentration of infective particles in the local
 ⁴⁵³ pool (Fig. 8c) and decreasing the particle uptake rate per individual (Fig. 8d)

454 3.3.2. Filtration/contact rate effect

455

To examine the importance of filtration or particle contact rate determining infection prevalence, we varied the 456 filtration/contact rate (Fig. 7b). The simulation showed that the lowest filtration or particle contact rate limits disease 457 transmission because, despite the incorporation of pathogens, the body burden remains below the infective dose 458 (Fig. 7b, gray line). For the two higher filtration rates, initially at day 100, an increasing filtration rate has a clear 459 effect on the prevalence of infection (20%). Beyond this point, the effect of filtration rate continuously decreases and, 460 eventually, beyond a certain point (ca day 700) where the prevalence of infection reaches a relatively steady maximum, 461 this effect is almost inappreciable (5%, day 700; 1%, day 1000). The eventual saturation of the local water volume 462 with pathogens (Fig. 6c) released by the increasing infected and dead populations (Fig. 3a,b) makes insignificant the 463 effect of filtration rate on prevalence as the disease progresses in the system. 464



Figure 7: Comparison of the effect of varying (a) recruitment rate, (b) the rate of filtration or contact with infective particles (c) the particle loss rate, and (d) the V_r : V_l ratio on prevalence of infection.

In addition to the contact rate of particles, for passive suspension feeders, such as corals, the accumulation of particles by the individual depends on the exposed surface area per individual A_s (see Eq. 2 and Fig. 11a). Considering Eq. (2), the importance of A_s determining disease prevalence would have the same pattern as the contact/filtration rate f_s . As the exposed surface area increases, more particles contact the animal, consequently prevalence of infection and the likelihood of an epizootic increases (see section 3.4). Similarly to f_s , the saturation of the water with pathogens may minimize the effect of the exposed surface area of the susceptible individual.

471 3.3.3. Waterborne pathogen loss

472

We examined the potential importance of infective particle loss influencing infection prevalence. For this purpose, 473 we varied the rate of particle loss in the local pool, that is the rate of inactivation of pathogens due to dilution, advec-474 tion, and mortality. Increasing particle loss limits the prevalence of infection (Fig. 7c) due to the reduced availability 475 of particles to be filtered or contacted. The progressive saturation of particles in the water volume accessible to the 476 benthic population (Fig. 6c) cancels the particle loss effect reducing the infection prevalence. Late in the simulation, 477 the release rate from infected animals clearly overwhelms the inactivation rate of pathogens in the water due to dilu-478 tion, advection, and natural mortality for the two lower particle loss scenarios tested (see solid and dashed black lines 479 in Fig. 7c). In contrast, at the highest particle loss rate tested (see gray solid line in Fig. 7c), the inactivation rate 480 of pathogens exceeds the release rate, resulting in the absence of infection for the first 800 days and a relatively low 481 prevalence (30%) in the last days of the simulation. 482

483 3.3.4. Diffusion of particles

484

Three rates of particle diffusion were simulated, by means of three remote volume V_r / local volume V_l ratios (Fig. 485 7d). For most suspension feeders, the volume directly influenced by filter or suspension feeding V_l will be small (e.g., 486 a volume with a height of 0-15 cm for oyster populations (Wilson-Ormond et al., 1997)). Consequently, the size of 487 the remote volume V_r may be a primary determinant of the prevalence of infection. While the local volume $(0.1 m^3)$ 488 was maintained constant, increasing the size of the remote volume from $0.1 m^3$ to $1 m^3$ resulted in a decrease in the 489 prevalence of infection of 40% at the beginning of the simulation. Initially, the two lower remote volumes tested 490 showed no infections (see black dashed and gray solid lines in Fig. 7d), but as the simulation progressed only the 491 larger remote volume limited disease transmission (see gray solid line in Fig. 7d). Similarly to the previous case, 492 where particle loss from the local pool was varied, the progressive saturation of particle concentration (Fig. 6a,b) 493 resulted, at the end of simulation, in a suppression of the diffusion effect initially observed for increasing remote 494 volumes. 495

496 3.4. Epizootiology and R_0

497

The generation of an epizootic is regulated by rates involved in the dynamics of both the host and the pathogen, 498 and factors associated with the particle diffusion processes. R_0 increases linearly with increasing transmission rate β_s , 499 and nonlinearly with increasing initial population N (Fig. 9a, black solid line), pathogen body burden and release rate 500 from infected animals $c_I b_I$ and dead infected animals $c_{DI} b_{DI}$, and with increasing filtration or contact rate of particles 501 f_s and local volume sizes V_l (Eq. (17)). The likelihood of an epizootic decreases linearly with increasing inactivation 502 rate of particles inside the body of the animal a, and nonlinearly with increasing disease mortality m_l , the removal of 503 dead individuals from the system d, the remote volume size V_r , the exchange rate of particles between local volume 504 and the remote volume γ , and the loss of particles in the local volume r and the remote volume σ (Eq. (16)). For 505 passive suspension feeders, in addition to filtration or contact rate of particles f_S , transmission of disease increases 506 nonlinearly with the surface area of the susceptible individual exposed to contact with waterborne pathogens A_s . 507

509 Host and non-focal host density

508

In a scenario where an infected animal is introduced into a totally susceptible population, once the population density rises sufficiently to compete for pathogens and lower the per capita body burden in the susceptible subpopulation, the probability of an epizootic developing remains relatively constant with increasing N (Fig. 9a, black solid line). In a scenario with relatively low pathogen body burden, this competition, more likely to occur in active filter feeders than in suspension feeders, results in a reduction of disease risk ($R_0 < 1$) (Fig. 9a, black dashed line).

In a system with non-focal suspension-feeding hosts H (Equation 20), with a relatively high filtration/contact rate of particles f_H , the competition effect between host types is intensified and, consequently, the likelihood of an epizootic decreases with increasing initial abundance of the non-focal host (Fig. 9a, gray solid line). Thus, the threshold of the initial host population N for an epizootic to occur may be substantially increased with increasing abundance of the non-focal host population. For instance, in the Fig. 9a scenario, the epizootic threshold for the initial host population in a system without non-focal filter feeders (N = 30) is increased to N = 80 after adding the same non-focal host

⁵¹⁰



Figure 8: Comparison of the effect of varying recruitment rate on (a) the susceptible population, (b) the infected population, (c) the particle concentration in the local volume, and (d) particle uptake rate.

⁵²² population (H = 30).

523

524 Remote volume size and particle loss

525

In a scenario with a relatively high exchange rate of particles between the local and remote pools ($\gamma = 1$), an increasing remote volume size, resulting in a diffusion-like transfer of particles to the remote pool, reduces the particle concentration in the local volume available for suspension-feeders and, thus, the likelihood of an epizootic (Fig. 9b, dashed line). The effect of increasing particle loss rate in the remote pool σ on disease development has a similar trend and intensity (Fig. 9b, dash-dot line). Note for comparison that the σ curve is for a remote volume size $V_r = 1$. The effect of increasing internal *in vivo* inactivation rate *a* also has a similar trend making R_0 estimate much more sensitive to small changes in this parameter (Fig. 9b).

533 534 Infective dose

535

Finally, we explore the similarity of increasing remote volumes and non-focal host densities, specifically focusing on the influence of the infective dose on disease development. When comparing the effect of increasing the remote volume size and non-focal host density on disease development for a high infective dose (200 particles), the R_0 estimate shows that an increase in non-focal host population density of 100 individuals m^{-2} has the same limiting effect



Figure 9: R_0 for increasing (a) initial susceptible N and non-focal host population density (individuals m^{-2}) with the overfiltration scenario for N (with half of the particles), and (b) remote volume size V_r , particle loss rate in the remote pool σ , and particle internal inactivation rate a. The black dotted line represents the outbreak threshold level $R_0 = 1$; above this level the likelihood of an epizootic is high; below this level, an outbreak is not expected.

on epizootics as a remote volume of $0.1 m^3$ (Fig. 10). This relationship is nonlinear; thus, a non-focal host population 540 of 500 individuals m^{-2} may have the same limiting effect on epizootics as a remote volume of 2.5 m^3 . For lower 541 infective doses (50 particles), the change in non-focal host population abundance or remote volume is much smaller 542 to obtain the same effect on reducing R_0 . For instance, an increase in non-focal host population from 0 to 200 individ-543 uals m^{-2} or in remote volume from 0 to 0.3 m^3 produces a drop in R_0 from 6 to 3, whereas for a high infective dose 544 (200 particles), the same increase in non-focal host population abundance or remote volume produces a much smaller 545 decrease (R_0 from 1.5 to 0.8). However, if the lower dose disease system has a higher R_0 than the high dose system, a 546 smaller drop in R_0 in the high dose system may contribute more importantly limiting an epizootic than the substantial 547 drop produced by the same change in non-focal host population abundance or remote volume in the low dose system 548 (Fig. 10). 549

550

⁵⁵¹ Passive suspension feeders and the exposed surface area

552

The transmission of disease in passive suspension-feeders not only depends on the contact rate with particles, but 553 also the surface area of the susceptible individual exposed to contact with waterborne pathogens (A_s) (Eq. (17)). First, 554 we conducted a simulation to evaluate the effect of varying A_s on the accumulation of particles. The simulation was 555 run with the conditions for parameters given in Table 2 for Case 2, with a carrying capacity K of 300 individuals. 556 Susceptible individuals accumulate particles (Fig. 11a) with time as disease progresses and the water becomes satu-557 rated with particles. The simulation shows an increase in the average host body burden of infective particles as the 558 exposed surface area increases. At the higher exposed surface area tested, the average internal number of particles per 559 susceptible increases to the infective dose (200 particles) by day 150. 560

Second, we introduced an infected animal in a susceptible population of 300 individuals and calculated R_0 for a range of values for of A_s , including possible realistic feeding surfaces in suspension feeders such as corals (Fig. 11b). These simulations were conducted for low and high infective doses (50 and 200 particles). The likelihood of an epizootic increases nonlinearly with the surface area exposed by the individual to waterborne pathogens A_s , resulting in the requirement for 3 times larger exposed surface area for an epizootic to occur in the higher infective dose case

566 (Fig. 11b).



Figure 10: R_0 for increasing non-focal host population density (individuals m^{-2}) and remote volume size (m^{-3}) for a high infective dose (200 infective particles) and for a low infective dose (50 particles). The gray dotted line represents the outbreak threshold level ($R_0 = 1$).

567 4. Discussion

568

The deterministic compartmental model accounts for the dynamics and epizootiology of waterborne microparasite infectious diseases in suspension-feeders, focusing on both the host and the pathogen dynamics. In this model, transmission occurs via particle uptake by contact for passive filter feeders or by filtration for active filter feeders, of waterborne infective pathogens under a body-burden based dose-response mechanism. The model yields a set the basic reproduction numbers R_0 which give insight about the relative importance of the parameters determining the outbreaks of epizootics. The generation of an epizootic is regulated by rates involved in the dynamics of both the host and the pathogen, and factors associated with the particle diffusion processes.

576 4.1. Host dynamics

577

Simulation results indicate that a relatively high initial concentration of infective particles available for suspension feeders, where particle inactivation is slow, the prevalence of disease increases rapidly in the first six months to a high and steady level (Fig. 4). Mortality is also relatively high, increasing to a 60% of the population at the end of the simulation. Disease mortality rates in bivalve filter feeders can be 50% or more per year. Rates this high are reported



Figure 11: Effect of increasing surface area of susceptible individuals exposed to contact with waterborne pathogens A_s on (a) particle accumulation with time, and (b) disease risk (R_0) for high (200 infective particles) and low (50 infective particles) infective doses. The gray dotted line represents the outbreak threshold level ($R_0 = 1$).

⁵⁸² in oysters in the Gulf of Mexico (Soniat and Brody, 1988). In Delaware Bay, the mortality rate is lower; however, dermo disease at least doubles the natural mortality rate of the market-size animals in epizootic years (Ford et al., 2006; Powell et al., 2008). The fungal disease Aspergillosis can impact sea fan corals with mortalities resulting in the loss of >50% of the sea fan colony area in six years (Kim and Harvell, 2004). Worldwide distributed clam species such as *Ruditpes philippinarum* show similar mortality rates associated to brown ring disease (Paillard et al., 2014).

The recruitment of new susceptible individuals to the system (Case 1, Fig. 4) permits levels of prevalence lower 587 than 100% (e.g., 90%, Fig. 4). This simulation with a low particle loss rate in the local volume of water saturated with 588 infective particles, likely reproduces a "water-tank" system where eventually all the particles in the water are coming 589 into contact with or filtered out by susceptible individuals and so prevalence in the population reaches high levels and 590 remains high, even with recruitment. The development and maintenance of high particle concentrations in the water 591 column (Fig. 6d) resulting in pandemic infection must be one of the reasons for the rapid transmission of Dermo 592 disease and the high mortalities in Eastern oyster populations observed in the early 1990s in Delaware Bay (Ford 593 et al., 2006; Bushek et al., 2012) and in other bays where most newly settled juveniles become infected well within 594 the first year of life (Powell et al., 1996; Ragone Calvo et al., 2003; McCollough et al., 2007; Soniat et al., 2012). 595 Geographically expansive regions with high concentrations of infective particles leading to pandemic infection must 596 occur commonly for diseases like Dermo and other Perkinsus-caused diseases in order to explain the continuously 597 high prevalence of infection (e.g., Park and Choi, 2001; Powell and Hofmann, 2015). 598

The number of dead animals is small ($<1 \text{ m}^{-2}$) in both simulations (Case 1 and 2) due to the high removal rate of these animals from the system. In general, in marine systems, the bacterial decomposition of organic matter (Allison,

1990; Lotz and Soto, 2002; Smith, 1953) or the action of scavengers removing dead animals (Veale et al., 2000; 601 Morello et al., 2005) is a relatively fast process and thus, dead animal tissue is removed quickly from the environment. 602 The initial increase of dead infected individuals to a maximum is related to the increase of infected individuals as the 603 disease transmission progresses in the population. The infected animals and their decay or scavenging can contribute 604 pivotally to transmission by the releasing of particles to the local volume and ultimately supplying the remote pool 605 (Fig. 6c) or they can be inconsequential if the act of decay or scavenging results in the loss of infectivity. Few studies 606 have examined the importance of post-mortem processes in disease transmission. 607

Simulations show the effect of varying transmission rate and carrying capacity on the host population (Case 1, 608 Fig. 3 vs Case 2, Fig. 4). When the population is below carrying capacity, the susceptible population initially 609 increases (Fig. 4) as the recruitment of new susceptible individuals is faster than the rate of infection. As the disease 610 transmission progresses in the population with a substantial increase of particles in the environment, the reduction 611 of susceptible individuals due to infection exceeds the incorporation of new individuals. The slower transmission 612 rate in Case 2 (Fig. 4) importantly reduces the spread of the disease in the population, lowering the prevalence of 613 infection. Overall, increasing transmission rate leads to higher prevalence of infection in filter feeders, especially 614 at low carrying capacities or at low population densities where the carrying capacity is high. Increasing population 615 density, and particularly in cases where carrying capacity is high, can reduce reduce prevalence of infection even for 616 relatively high disease transmission rates (Fig. 5). In a contact-based density-dependent disease, a population at its 617 carrying capacity has a higher likelihood of disease transmission than when the population is below carrying capacity 618 (Gao and Hethcote, 1992). However, dense filter feeder populations at carrying capacity may be less vulnerable to 619 disease; individual hosts may 'compete' for pathogens reducing the concentration of infective particles sufficiently to 620 limit body burden below the infective dose and, in turn, limiting the probability of epizootic development (i.e. the

overfiltration scenario, Bidegain et al., in press). 622

4.2. Particle dynamics 623

624

621

Many disease models for marine filter-feeders have addressed the proliferation of infection and subsequent disease 625 impact, accepting widespread and rapid transmission to be the norm (Ford et al., 1999; Powell et al., 1999, 1996). 626 Others have examined the effects of diseases on host population dynamics (Kuris and Lafferty, 1992; Yakob and 627 Mumby, 2011). However, models of marine diseases that incorporate particles as a variable, permitting the water 628 column to act as a 'reservoir', conduit, or sink for infective particles, are few and none also include an infective 629 dose effect based on an in vivo compartment of infective particles sequestered in the susceptible pool of hosts. The 630 model that we propose here tracks pathogen dynamics in (i) the local volume with a particle pool interacting with the 631 host population, (ii) a remote volume that can act as a source or sink for infective particles depending on the relative 632 concentration between the two volumes, and (iii) a transient pool of infective particles in the susceptible individuals 633 that may ultimately be inactivated or transition the individual to the infected pool. 634

Simulations show that initial infections of a few susceptible individuals initiates the process (Fig. 3a,b) by which 635 substantial numbers of infective particles are released into the water thereby becoming available to susceptibles and 636 saturating the water with particles (Fig. 6c). This results in an initial diffusion-like transfer of particles to the adjacent 637 remote volume and, eventually, in a 'balance' in particle concentration between the two water compartments (Fig. 6d). 638 A system with a high release rate of pathogens from infected animals or a high rate of release of infective particles upon 639 death, and a limited inactivation rate of infective particles either in the water column or in the susceptible host, might 640 easily lead to be a highly transmissible system characterized by a relatively high incidence of disease outbreaks. This 641 is the case, for instance, for oysters and the pathogen Perkinsus marinus (Bushek et al., 2002), but might also occur in 642 corals, where the decay of tissue of infected colonies infected by black-band disease or aspergillosis releases infectious 643 fomites or individual particles into the water which then drift to nearby corals (Jolles et al., 2002; Richardson, 2004; 644 Zvuloni et al., 2009) or in the case of withering syndrome in abalone where infective elements or fomites released 645 by infected individuals before or after death spread infection over wide expanses (Lafferty et al., 2015). We suspect 646 a similar scenario for the disease of long-spined sea urchins that ravaged the Caribbean populations late in the last 647 century (Lessios et al., 1984). Consequently, in these systems, parameters incorporated into the model, such as the 648 body burden of pathogens in infected (b_I) or dead animals (b_{DI}) and the relative importance of the release rate (c) 649 and the removal rate of pathogens in the tissue (a) or water column (r and sigma) might have profound influence in 650 determining the frequency and geographic extensiveness of epizootics. 651

The model present here also tracks particle dynamics inside the susceptible population. The dynamics of the 652 infective dose effect in filter and suspension feeders is poorly understood (e.g., Chu and Volety, 1997; Wang et al., 653 2010). Our model suggest that the accumulation of particles increases to a maximum in the initial phase of the disease 654 process as susceptible animals take up particles without reaching an infective dose and that a number of *in vivo* and 655 external processes determine the degree of this accumulation and the probability that sufficient infective particles 656 will be accumulated to produce an infection. Among these is the immune response, but equally important are the 657 processes controlling uptake, namely the particle acquisition rate (filtering or surface area impingement) and particle 658 concentration which itself is controlled by the rate of addition and loss of particles from the local water volume. This 659 dynamic is likely enormously important, yet to date has been little studied. Assuming that the disease transmission in 660 filter feeders (Bushek et al., 1997; Ford et al., 1999; Powell et al., 1999), and probably in other suspension feeders, 661 occurs via an infective dose so that susceptible animals need to have some minimum level of body burden of infectious 662 particles before they become infected, some susceptible individuals will then have a relatively high body burden while 663 most will have a small body burden (see section 2.4.1). Presumably, these animals will have some unique attributes 664 either genetically or environmentally that has resulted in their higher body burdens. Understanding the contributing 665 factors that result in a few animals initiating infections than then can propagate by supporting increased particle 666 concentration in the water column is a critical need. In this regard, simulations show that the average per capita 667 dose in susceptible population is 150 particles (Fig. 6b). This demonstrate the adequate performance of dose-based 668 transmission term, considering that the average per capita dose is lower than the minimum dose for susceptibles to 669 become infected b_{min} (200 particles). 670

671 4.3. Recruitment

672

Recruitment introduces new, healthy individuals into the population which have the net effect of reducing disease 673 prevalence. However, in this case, this effect of recruitment on reducing prevalence does not directly limit disease; a 674 decrease in prevalence of infection is not a consequence, for instance, of new individuals increasing population particle 675 filtration rate and then reducing the local pool of infective particles sufficiently to decrease per capita accumulation 676 below the infective dose. The prevalence is lower only due to the addition of more susceptibles to the system. The 677 influence of these new susceptible individuals will depend upon their role as new produces of infective particles or as 678 new 'inactivators' of infective particles. Often these new recruits will promote the disease as the number of individuals 679 becoming infected increases (Fig. 8b) despite the reduction of (i) particles in the water column due to increasing 680 population particle filtration (Fig 8c) and (ii) the particle uptake rate per animal (Fig. 8d). Rather, these individuals 681 provide increased production of infective particles, raising their concentration in the local pool and ultimately the 682 remote pool. This explains the occurrence of pandemic infections in marine filter feeders such as oysters regardless 683 of increasing recruitment intensities (Ford, 1996; Bushek et al., 2012). 684

However, the effect of recruitment in limiting the occurrence or the duration of disease epizootics can be substantial in both active and passive suspension feeders when enough recruits enter the population to reduce the infective particle concentration in the water column such that the particle uptake rate per animal drops to a point where an increasing number of animals retain a transient body burden below the infective dose. This effect has been suggested to occur in oyster populations following high intensity recruitment events. Even at high salinity, suitable for disease transmission, an oyster population might be able to resist epizootic development as long as recruitment sustains the increase in population density (Hofmann et al., 1995; Powell et al., 1996).

692 4.4. Filtration/contact rate

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Filtration rate in filter feeders, such as bivalves, can be affected by diverse characteristics of the particles filtered, such as the concentration of phytoplankton and suspended solids and the quality and size of food particles (Hofmann et al., 1994; Khalil, 1996). The physical parameters of the natural habitat such as temperature, salinity, and water flow (MacDonald and Thompson, 1986) and the size of the animal can also affect the filtration rate (Dame, 2011). In passive suspension feeders, such as corals, the contact rate depends on the polyp or colony exposed surface area and the flow conditions (Sebens et al., 1996, 1998). Such factors affect the incorporation of pathogens by susceptible

⁷⁰⁰ animals and, thus, the disease incidence.

We varied filtration/contact rates to explore this effect (Fig. 7b). A relatively low filtration or particle contact 701 rate limits disease transmission, due to lower per capita exposure to pathogens. Increasing filtration rate has an effect 702 on disease transmission when the concentration of particles in the water is relatively low. In this case, as filtration 703 or contact rate increases, the rate of disease transmission increases correspondingly. However, a change in filtration 704 or contact rate ceases to exert a significance influence on transmission when the local volume becomes saturated 705 with infective particles. At this point, higher filtration rates (Fig. 7b) result in the same high prevalence; that is, the 706 per capita dose received by individuals is above the infectious dose over a wide range of population filtration rates. 707 However, active filter feeders in particular, and suspension feeders possibly have two mechanisms, explored here, 708 to reduce the incidence of infection despite having high filtration rates and contact rates with substantive numbers 709 of infective particles in the water column. The first obviously is a higher infective dose which might result from 710 selection for more disease resistant genotypes, for example (Munroe et al., 2015) (Fig. 10) and the second is a dense 711 assemblage of hosts competing for pathogens and reducing the per capita dose (i.e. the overfiltration scenario -712 Bidegain et al. in press) (Fig. 9a, black dashed line). Because it is well established that dense concentrations of 713 bivalves can become food limited (Heasman et al., 1998; Jiang and Gibbs, 2005; Powell et al., 1995), it stands to 714 reason that the same process would limit per-capita incorporation rates of infective particles. In passive suspension 715 feeder hosts in environments with high particle concentrations and contact rates, reduced exposed feeding areas may 716 result in lowering the accumulation of particles (Fig. 11a), and thus the prevalence of infection. Whether passive 717 suspension feeders can become sufficiently dense to limit per capita particle uptake rate is unclear. 718

719 4.5. Particles loss and diffusion

720

The model incorporates two parameters for particle loss in the water column. The parameter r accounts for the 721 particle loss within the local volume V_l due to natural mortality or other processes of inactivation such as sedimenta-722 tion, and advection and diffusion to the remote volume V_r . The parameter σ accounts for the subsequent inactivation 723 through natural mortality or other sources of inactivation such as sedimentation in this remote pool or perhaps loss 724 from this compartment through, for example, outwelling. These two volumes in the model are formulated in order to 725 account for particle diffusion by means of concentration gradients between them. Initially, in the simulation when the 726 local volume is not saturated with particles, increasing particle loss through diffusion of particles to the remote pool 727 can limit epizootic development. When, the environment is saturated with particles, this effect becomes inconsequen-728 tial (Fig. 7c,d) unless the inactivation rate in the remote pool is high (Fig. 9b). A large remote volume together with a 729 high exchange rate of particles between pools γ and a relatively high inactivation rate of pathogens in the remote pool 730 a is an effective mechanism to reduce particle concentration locally and prevent transmission (Bidegain et al., in press) 731 (Fig. 9b). Thus, a scenario similar to what is shown in Fig. 7d (gray line), in which the remote volume is substantially 732 larger than the local volume, thereby leading to a rapid transfer of pathogens from the local pool to the remote pool, 733 is particularly effective at limiting the incidence of infection. This scenario could easily occur in estuaries with short 734 water residence times (Armstrong, 1982; Marshall and Alden, 1997)during periods of increased freshwater inflow or 735 during spring tides where the larger tidal volume results in increased outwelling of particles, thereby reducing particle 736 concentration in the water column (Ellien et al., 2004; Banas et al., 2007; Gray et al., 2009; Narváez et al., 2012). 737

On the other hand, a high concentration of infective particles in the remote volume could overcome any local loss 738 mechanisms such as overfiltration or the removal by non-focal hosts, and thus maintain high prevalence. The dynamic 739 interchange between local and remote pools if not unique to the marine realm is at least particularly characteristic of 740 the disease transmission processes taking place there. This model is a theoretical model that explores the processes 741 that initiate and terminate an epizootic. The simplification of hydrodynamics to two volumes, the volume directly 742 impacted by the filter feeder and the volume overlying it, with particle diffusion occurring across the interface is 743 sufficient to explain these processes theoretically. The study of a specific case such as the disease transmission on 744 a reef or in an estuary would require the use of similar disease models coupled to a hydrodynamic model. While 745 complex models of larval dispersal using location-specific hydrodynamics have been employed often (North et al., 746 2008; Ayata et al., 2009; Bidegain et al., 2013), the transmission of disease particles in such applications has been 747 little investigated (Murray and Jackson, 1992). 748

749 4.6. Epizootiology and R_0

⁷⁵¹ Simulation results of host and particle dynamics showed that dead infected animals, likely with higher body ⁷⁵² burden and release rates of infective particles than most live infected animals (Bushek et al., 2002; Richardson, 2004), ⁷⁵³ are determinant in the generation of epizootics in benthic filter and suspension feeders. A relatively small external ⁷⁵⁴ source of infective particles initially available to a benthic suspension–feeder population may infect few individuals. ⁷⁵⁵ Although this external source is consumed rapidly or lost and inactivated in the water column, these initially infected ⁷⁵⁶ individuals can trigger the generation of the epizootic by means of particle release in the live-infected stage and, ⁷⁵⁷ particularly, upon death.

However, mechanisms potentially exist that might lead to a lowering of the epizootic likelihood in a benthic 758 population of suspension feeders even if a relatively large number of particles is available. High densities of sus-759 pension feeders, particularly filter-feeders, may compete for food, but also for waterborne pathogens. Moreover, a 760 non-focal host population ingesting or otherwise collecting infective particles (e.g., bivalves, sponges or tunicates), 761 may enhance the competition effect, which could lead to a lower per capita exposure to pathogens (Peterson and 762 Andre, 1980; Beukema and Cadée, 1996; Powell et al., 2012c). Thus, the threshold of the initial host population 763 for an epizootic to occur may be substantially increased with increasing numbers of non-focal hosts (Fig. 9). This 764 intra- or inter-specific competition (Fig. 9a and 8) by itself or together with particle diffusion (Fig. 9b) may reduce 765 substantially the likelihood of an epizootic. 766

The effect of increasing particle loss on limiting the epizootic development is similar to that produced by an 767 increase in remote volume size (Fig. 9b). This means that the remote volume is an effective mechanism for limiting 768 an epizootic as it can remove particles from the local pool by diffusion. Thus, an increase in the remote volume size 769 may exert a limiting effect on epizootic development comparable to that produced by inter-specific competition with 770 a non-focal host population. Moreover, an increase in these two disease limiting factors will have a higher impact on 771 disease control in hosts with a relatively high infective dose (Fig. 10). However, an increase in the remote volume 772 needs to be accompanied by particle inactivation or loss (i.e. σ) in order to have an effect on epizootic development. 773 In absence of this particle removal, the remote volume may become saturated with particles and diffuse them back to 774 the local pool thereby making them available to susceptible hosts. In contrast, in vivo inactivation rate of particles, 775 by limiting particle body burden below the infective dose, has a more intense effect on disease development than the 776 remote volume size or the particle loss in this volume (Fig. 9b). 777 Overall, these disease limiting mechanisms apply similarly, but no in equivalent measure, to active and passive 778

suspension feeders. In the case of passive suspension feeders, the competition for particles may be less effective than 779 in active filter-feeders. The exposed surface area of the individuals A_s in a colony of passive suspension feeders (e.g. 780 corals) is as important as the density of individuals; thus, high density populations with a small exposed surface area 781 per individual, resulting in a reduced particle accumulation, may limit the disease outbreak (Fig. 11) just as well as 782 fewer individuals with a large surface area for particle capture. Thus, it seems that the active foraging strategy of filter 783 feeders permits these organisms to effectively limit disease outbreaks in comparison to passive feeders, but the same 784 attribute makes them more vulnerable to epizootics when population densities are inadequate to control local particle 785 concentrations beneath the threshold yielding an effective dose (Fig. 9a and Fig. 11). 786

787 5. Conclusions

788

A large number of marine diseases are transmitted through the water column from one host to another. This 789 water column provides a 'reservoir' for infective particles and the mechanisms by which particles are added to it or 790 lost from it exert an important influence on the prevalence of disease and more importantly the difference between a 791 disease exerting a local impact on a host population and pandemic disease affecting the host over large geographic 792 regions. The local population modulates this effect through genetic characteristics that affect the infective dose and 793 through varying local availability by modulating particle incorporation and release rates. The dynamic imposed by 794 this complex interaction between population and water column, potentially over metapopulation scales, is relatively 795 unique to the marine world. Understanding the details of this dynamic is critical to understanding the disease process 796 in host populations and to improving management responses to marine disease challenges. Models that incorporate 797 these processes, such as the one described here, are important tools, little used heretofore in the marine context, to 798 explain the epizootiology of marine diseases. Only when a disease can be understood at the *in vivo* scale of the 799

⁸⁰⁰ individual, the local scale of the population, and the geographic scale of the metapopulation will effective approaches ⁸⁰¹ to management become routinely achievable.

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805 References

- Allison, P., 1990. Variation in rates of decay and disarticulation of Echinodermata: implications for the application of actualistic data. Palaios 5, 432–440.
- Anderson, R., Gordon, D., 1982. Processes influencing the distribution of parasite numbers within host populations with special emphasis on parasite-induced host mortalities. Parasitology 85, 373–398.
- Anderson, R. M., May, R. M., 1981. The population dynamics of microparasites and their invertebrate hosts. Philosophical Transactions of the
 Royal Society of London. Series B, Biological Sciences 291, 451–524.
- Armstrong, N. E., 1982. Response of Texas estuaries to freshwater inflows. In: Kennedy, V. S. (Ed.), Proc. 6th Biennial International Estuarine
 Research Conference. Academic Press, pp. 103–120.
- Audemard, C., Ragone Calvo, L., Paynter, K., Reece, K., Burreson, E., 2006. Real-time PCR investigation of parasite ecology: *in situ* determination
 of oyster parasite *Perkinsus marinus* transmission dynamics in lower Chesapeake Bay. Parasitology 132, 827–842.
- Ayata, S.-D., Ellien, C., Dumas, F., Dubois, S., Thiébaut, É., 2009. Modelling larval dispersal and settlement of the reef-building polychaete
 Sabellaria alveolata: role of hydroclimatic processes on the sustainability of biogenic reefs. Continental Shelf Research 29, 1605–1623.
- Bak, R., Joenje, M., De Jong, I., Lambrechts, D., Nieuwland, G., 1998. Bacterial suspension feeding by coral reef benthic organisms. Marine
 Ecology Progress Series 175, 285–288.
- Banas, N. S., Hickey, B. M., Newton, J. A., Ruesink, J. L., 2007. Tidal exchange, bivalve grazing, and patterns of primary production in Willapa
 Bay, Washington, USA. Marine Ecology Progress Series 341, 123–139.
- Beukema, J. J., Cadée, G. C., 1996. Consequences of the sudden removal of nearly all mussels and cockles from the Dutch Wadden Sea. Marine Ecology 17, 279–289.
- Bidegain, G., Bárcena, J. F., García, A., Juanes, J. A., 2013. LARVAHS: predicting clam larval dispersal and recruitment using habitat suitability based particle tracking model. Ecological Modelling 268, 78–92.
- Bidegain, G., Powell, E. E., Klinck, J. M., Ben-Horin, T., Hofmann, E. E., 2016. Marine infectious disease dynamics and outbreak thresholds:
 contact, transmission, pandemic infection, and the potential role of filter feeders. Ecosphere, in press.
- Bushek, D., Ford, S. E., Alcox, K. A., Gustafson, R., Allen Jr., S. K., 1997. Response of the eastern oyster, *Crassostrea virginica* to *in vitro* cultured
 Perkinsus marinus and the early fate of parasites delivered via three dosing methods. Journal of Shellfish Research 16, 479–485.
- Bushek, D., Ford, S. E., Burt, I., 2012. Long-term patterns of an estuarine pathogen along a salinity gradient. Journal of Marine Research 70,
 225–251.
- Bushek, D., Ford, S. E., Chintala, M. M., 2002. Comparison of *in vitro*-cultured and wild-type *Perkinsus marinus*. III. Fecal elimination and its role
 in transmission. Diseases of Aquatic Organisms 51, 217–225.
- Calvo, L. M. R., Wetzel, R. L., Burreson, E. M., 2001. Development and verification of a model for the population dynamics of the protistan
 parasite, *Perkinsus marinus*, within its host, the eastern oyster, *Crassostrea virginica*, in Chesapeake Bay. Journal of Shellfish Research 20,
 231–241.
- Choi, K.-S., Powell, E. N., Lewis, D. H., Ray, S. M., 1994. Instantaneous reproductive effort in female American oysters, *Crassostrea virginica*,
 measured by a new immunoprecipitation assay. The Biological Bulletin 186, 41–61.
- Choi, K.-S., Wilson, E. A., Lewis, D. H., Powell, E. N., Ray, S. M., 1989. The energetic cost of *Perkinsus marinus* parasitism in oysters: quantification of the thioglycollate method. Journal of Shellfish Research 8, 125–131.
- Chu, F.-L. E., Volety, A. K., 1997. Disease processes of the parasite *Perkinsus marinus* in eastern oyster *Crassostrea virginica*: minimum dose for
 infection initiation, and interaction of temperature, salinity and infective cell dose. Diseases of Aquatic Organisms 28, 61–68.
- ⁸⁴³ Cuddington, K., Beisner, B., 2005. Ecological paradigms lost: routes of theory change. Theoretical Ecology Series. Elsevier Academic, Burlington,
 ⁸⁴⁴ Massachusetts.
- Dame, R. F., 1993. The role of bivalve filter feeder material fluxes in estuarine ecosystems. In: Bivalve filter feeders, in Estuarine and Coastal
 Ecosystem Processes. Vol. 33 of NATO ASI Series. Springer-Verlag, Berlin, pp. 245–269.
- Dame, R. F., 2011. Ecology of marine bivalves: an ecosystem approach. CRC Marine Science. CRC Press, Boca Raton, FL.
- Dang, C., de Montaudouin, X., Caill-Milly, N., Trumbic, Z., 2010. Spatio-temporal patterns of perkinsosis in the Manila clam *Ruditapes philippinarum* from Arcachon Bay (SW France). Diseases of Aquatic Organisms 91, 151–159.
- Diamond, E. A., 2012. Do scavengers influence dermo disease (*Perkinsus marinus*) transmission?: experiments in oyster parasite trophic interac tions. MSc Thesis, Rutgers University-Graduate School-New Brunswick.
- ⁸⁵² Diekmann, O., Heesterbeek, H., Britton, T., 2013. Mathematical tools for understanding infectious disease dynamics. Princeton University Press,
 ⁸⁵³ Woodstock, Oxfordshire.
- ⁸⁵⁴ Diekmann, O., Heesterbeek, J. A. P., Metz, J. A. J., 1990. On the definition and the computation of the basic reproduction ratio R_0 in models for ⁸⁵⁵ infectious diseases in heterogeneous populations. Journal of Mathematical Biology 28, 365–382.
- ⁸⁵⁶ Diekmann, O., Heesterbeek, J. A. P., Roberts, M. G., 2010. The construction of next-generation matrices for compartmental epidemic models.
 ⁸⁵⁷ Journal of the Royal Society Interface 7, 873–885.

- ⁸⁵⁸ Dietz, K., 1993. The estimation of the basic reproduction number for infectious diseases. Statistical Methods in Medical Research 2, 23–41.
- Ellien, C., Thiébaut, E., Dumas, F., Salomon, J.-C., Nival, P., 2004. A modelling study of the respective role of hydrodynamic processes and larval
 mortality on larval dispersal and recruitment of benthic invertebrates: example of *Pectinaria koreni* (Annelida: Polychaeta) in the Bay of Seine
 (English Channel). Journal of Plankton Research 26, 117–132.
- Ertman, S. C., Jumars, P. A., 1988. Effects of bivalve siphonal currents on the settlement of inert particles and larvae. Journal of Marine Research 46, 797–813.
- Ford, S. E., 1992. Avoiding the transmission of disease in commercial culture of molluscs, with special reference to *Perkinsus marinus* (Dermo)
 and *Haplosporidium nelsoni* (MSX). Journal of Shellfish Research 11, 539–546.
- Ford, S. E., 1996. Range extension by the oyster parasite *Perkinsus marinus* into the northeastern United States: response to climate change. Journal
 of Shellfish Research 15, 45–56.
- Ford, S. E., Cummings, M. J., Powell, E. N., 2006. Estimating mortality in natural assemblages of oysters. Estuaries and Coasts 29, 361–374.
- Ford, S. E., Schotthoefer, A., Spruck, C., 1999. *In vivo* dynamics of the microparasite *Perkinsus marinus* during progression and regression of
 infections in Eastern oysters. The Journal of Parasitology 85, 273–282.
- Ford, S. E., Smolowitz, R., 2007. Infection dynamics of an oyster parasite in its newly expanded range. Marine Biology 151, 119–133.
- Ford, S. E., Tripp, M. R., 1996. Diseases and defense mechanisms. In: Kennedy, V. S., Newell, R. I. E., Eble, A. E. (Eds.), The Eastern Oyster,
 Crassostrea Virginica. Maryland Sea Grant College Park, pp. 581–660.
- Fréchette, M., Aitken, A. E., Page, L., 1992. Interdependence of food and space limitation of a benthic suspension feeder: consequences for
 self-thinning relationships. Marine Ecology Progress Series 83, 55–62.
- 676 Gao, L. Q., Hethcote, H. W., 1992. Disease transmission models with density-dependent demographics. Journal of Mathematical Biology 30, 717–731.
- Gedan, K. B., Kellogg, L., Breitburg, D. L., 2014. Accounting for multiple foundation species in oyster reef restoration benefits. Restoration
 Ecology 22, 517–524.
- Gilmour, J. P., Smith, L. D., Heyward, A. J., Baird, A. H., Pratchett, M. S., 2013. Recovery of an isolated coral reef system following severe disturbance. Science 340, 69–71.
- Gray, B. R., Bushek, D., Drane, J. W., Porter, D., 2009. Associations between land use and *Perkinsus marinus* infection of eastern oysters in a high
 salinity, partially urbanized estuary. Ecotoxicology 18, 259–269.
- Heasman, K. G., Pitcher, G. C., McQuaid, C. D., Hecht, T., 1998. Shellfish mariculture in the Benguela system: raft culture of *Mytilus galloprovin- cialis* and the effect of rope spacing on food extraction, growth rate, production, and condition of mussels. Journal of Shellfish Research 17,
 33–40.
- Hoese, H. D., 1962. Studies on oyster scavengers and their relation to the fungus *Dermocystidium marinum*. Proceedings of the National Shellfisheries Association 53, 161–174.
- Hofmann, E. E., Klinck, J. M., Powell, E. N., Boyles, S., Ellis, M., 1994. Modeling oyster populations II. Adult size and reproductive effort. Journal
 of Shellfish Research 13, 165–182.
- Hofmann, E. E., Powell, E. N., Klinck, J. M., Saunders, G., 1995. Modelling diseased oyster populations I. Modelling *Perkinsus marinus* infections
 in oysters. Journal of Shellfish Research 14, 121–151.
- Jiang, W., Gibbs, M. T., 2005. Predicting the carrying capacity of bivalve shellfish culture using a steady, linear food web model. Aquaculture 244, 171–185.
- Jolles, A. E., Sullivan, P., Alker, A. P., Harvell, C. D., 2002. Disease transmission of *Aspergillosis* in sea fans: inferring process from spatial pattern. Ecology 83, 2373–2378.
- Kermack, W. O., McKendrick, A. G., 1927. A contribution to the mathematical theory of epidemics. Proceedings of the Royal Society of London.
 Series A, Containing Papers of a Mathematical and Physical Character 115, 700–721.
- Khalil, A., 1996. The influence of algal concentration and body size on filtration and ingestion rates of the clam *Tapes decussatus*. Aquaculture
 Research 27, 613–621.
- 1901 Kim, K., Harvell, C. D., 2004. The rise and fall of a six-year coral-fungal epizootic. The American Naturalistmerican naturalist 164, 52–63.
- Kuris, A. M., Lafferty, K. D., 1992. Modelling crustacean fisheries: effects of parasites on management strategies. Canadian Journal of Fisheries
 and Aquatic Sciences 49, 327–336.
- Lafferty, K. D., Harvell, C. D., Conrad, J. M., Friedman, C. S., Kent, M. L., Kuris, A. M., Powell, E. N., Rondeau, D., Saksida, S. M., 2015.
 Infectious diseases affect marine fisheries and aquaculture economics. Annual Review of Marine Science 7, 471–496.
- Lessios, H., Cubit, J., Robertson, D., Shulman, M., Parker, M., Garrity, S., Levings, S., 1984. Mass mortality of *Diadema antillarum* on the Caribbean coast of Panama. Coral Reefs 3, 173–182.
- Liao, C.-M., Yeh, C.-H., Chen, S.-C., 2008. Predation affects the susceptibility of hard clam *Meretrix lusoria* to Hg-stressed birnavirus. ecological modelling 210, 253–262.
- Lotz, J. M., Soto, M. A., 2002. Model of white spot syndrome virus (WSSV) epidemics in *Litopenaeus vannamei*. Diseases of Aquatic Organisms
 50, 199–209.
- MacDonald, B., Thompson, R., 1986. Influence of temperature and food availability on the ecological energetics of the giant scallop *Placopecten magellanicus*. Marine Biology 93, 37–48.
- Marshall, H. G., Alden, R. W., 1997. Dynamics of an estuarine ecosystem: the influence of flow patterns on phytoplankton trends in the Chesapeake
 Bay. Oceanologica Acta 20, 109–117.
- McCallum, H., Gerber, L., Jani, A., 2005. Does infectious disease influence the efficacy of marine protected areas? A theoretical framework.
 Journal of Applied Ecology 42, 688–698.
- McCallum, H., Harvell, D., Dobson, A., 2003. Rates of spread of marine pathogens. Ecology Letters 6, 1062–1067.
- McCallum, H. I., Kuris, A., Harvell, C. D., Lafferty, K. D., Smith, G. W., Porter, J., 2004. Does terrestrial epidemiology apply to marine systems?
 Trends in Ecology and Evolution 19, 585–591.
- McCollough, C. B., Albright, B. W., Abbe, G. R., Barker, L. S., Dungan, C. F., 2007. Acquisition and progression of perkinsus marinus infections by specific-pathogen-free juvenile oysters (crassostrea virginica gmelin) in a mesohaline chesapeake bay tributary. Journal of Shellfish Research

923 26, 465-477.

Monismith, S. G., Koseff, J. R., Thompson, J. K., O'Riordan, C. A., Nepf, H. M., 1990. A study of model bivalve siphonal currents. Limnology and Oceanography 35, 680–696.

Morello, E. B., Froglia, C., Atkinson, R. J. A., Moore, P. G., 2005. Impacts of hydraulic dredging on a macrobenthic community of the Adriatic Sea, Italy. Canadian Journal of Fisheries and Aquatic Sciences 62, 2076–2087.

Munroe, D. M., Powell, E. N., Ford, S. E., Hofmann, E. E., Klinck, J. M., 2015. Outcomes of asymmetric selection pressure and larval dispersal
 on evolution of disease resistance: a metapopulation modeling study with oysters. Marine Ecology Progress Series 531, 221–239.

Murray, A. G., Jackson, G. A., 1992. Viral dynamics: A model of the effects size, shape, motion and abundance of single-celled planktonic organisms and other particles. Marine Ecology Progress Series. Oldendorf 89, 103–116.

Narváez, D. A., Klinck, J. M., Powell, E. N., Hofmann, E. E., Wilkin, J., Haidvogel, D. B., 2012. Modeling the dispersal of eastern oyster
 (*Crassostrea virginica*) larvae in Delaware Bay. Journal of Marine Research 70, 381–409.

- Newell, R. I., 2004. Ecosystem influences of natural and cultivated populations of suspension-feeding bivalve molluscs: a review. Journal of Shellfish Research 23, 51–62.
- North, E. W., Schlag, Z., Hood, R., Li, M., Zhong, L., Gross, T., Kennedy, V. S., 2008. Vertical swimming behavior influences the dispersal of
 simulated oyster larvae in a coupled particle-tracking and hydrodynamic model of Chesapeake Bay. Marine Ecology Progress Series 359, 99.
- Ostfeld, R. S., Keesing, F., 2000. Biodiversity series: the function of biodiversity in the ecology of vector-borne zoonotic diseases. Canadian Journal
 of Zoology 78, 2061–2078.
- Paillard, C., 2004. A short-review of brown ring disease, a vibriosis affecting clams, *Ruditapes philippinarum* and *Ruditapes decussatus*. Aquatic
 Living Resources 17, 467–475.
- Paillard, C., Jean, F., Ford, S. E., Powell, E. N., Klinck, J. M., Hofmann, E. E., Flye-Sainte-Marie, J., 2014. A theoretical individual-based model
 of brown ring disease in Manila clams textitVenerupis philippinarum. Journal of Sea Research 91, 15–34.
- Park, K.-I., Choi, K.-S., 2001. Spatial distribution of the protozoan parasite perkinsus sp. found in the manila clams, ruditapes philippinarum, in
 korea. Aquaculture 203, 9–22.
- Peterson, C. H., Andre, S. V., 1980. An experimental analysis of interspecific competition among marine filter feeders in a soft-sediment environ ment. Ecology, 129–139.
- Porter, E. T., Cornwell, J. C., Sanford, L. P., Newell, R., 2004. Effect of oysters *Crassostrea virginica* and bottom shear velocity on benthic-pelagic
 coupling and estuarine water quality. Marine Ecology Progress Series 271, 61–75.
- Powell, E. N., Ashton-Alcox, K. A., Kraeuter, J. N., Ford, S. E., Bushek, D., 2008. Long-term trends in oyster population dynamics in Delaware
 Bay: regime shifts and response to disease. Journal of Shellfish Research 27, 729–755.
- Powell, E. N., Hofmann, E. E., 2015. Models of marine molluscan diseases: Trends and challenges. Journal of Invertebrate Pathology 131, 212–225.
- Powell, E. N., Hofmann, E. E., Klinck, J. M., 1996. Modeling diseased oyster populations II. Triggering mechanisms for *Perkinsus marinus* epizootics. Journal of Shellfish Research 15, 141–165.
- Powell, E. N., Hofmann, E. E., Klinck, J. M., Ray, S. M., 1992. Modeling oyster populations: I. A commentary on filtration rate. Is faster always
 better? Journal of Shellfish Research 11, 387–398.
- Powell, E. N., Klinck, J. M., Ashton-Alcox, K., Hofmann, E. E., Morson, J., 2012a. The rise and fall of *Crassostrea virginica* oyster reefs: the role
 of disease and fishing in their demise and a vignette on their management. Journal of Marine Research 70, 505–558.
- Powell, E. N., Klinck, J. M., Ashton-Alcox, K. A., Kraeuter, J. N., 2009. Multiple stable reference points in oyster populations: biological
 relationships for the eastern oyster (*Crassostrea virginica*) in delaware bay. Fishery Bulletin 107, 109–132.
- Powell, E. N., Klinck, J. M., Ford, S. E., Hofmann, E. E., Jordan, S. J., 1999. Modeling the MSX parasite in eastern oyster (*Crassostrea virginica*)
 populations. III. Regional application and the problem of transmission. Journal of Shellfish Research 18, 517–538.
- Powell, E. N., Klinck, J. M., Guo, X., Ford, S. E., Bushek, D., 2011. The potential for oysters, *Crassostrea virginica*, to develop resistance to
 Dermo disease in the field: evaluation using a gene-based population dynamics model. Journal of Shellfish Research 30, 685–712.
- Powell, E. N., Klinck, J. M., Guo, X., Hofmann, E. E., Ford, S. E., Bushek, D., 2012b. Can oysters *Crassostrea virginica* develop resistance to dermo disease in the field: the impediment posed by climate cycles. Journal of Marine Research 70, 309–355.
- Powell, E. N., Klinck, J. M., Hofmann, E. E., Wilson-Ormond, E. A., Ellis, M. S., 1995. Modeling oyster populations. v. declining phytoplankton stocks and the population dynamics of American oyster (*Crassostrea virginica*) populations. Fisheries Research 24 (3), 199–222.
- Powell, E. N., Kreeger, D. A., Morson, J. M., Haidvogel, D. B., Wang, Z., Thomas, R., Gius, J. E., 2012c. Oyster food supply in Delaware Bay:
 estimation from a hydrodynamic model and interaction with the oyster population. Journal of Marine Research 70, 469–503.
- Powell, E. N., White, M., Wilson, E., Ray, S., 1987. Small-scale spatial distribution of oysters (*Crassostrea virginica*) on oyster reefs. Bulletin of
 Marine Science 41, 835–855.
- Ragone Calvo, L. M., Calvo, G. W., Burreson, E. M., 2003. Dual disease resistance in a selectively bred eastern oyster *Crassostrea virginica*, strain
 tested in Chesapeake Bay. Aquaculture 220, 69–87.
- Regoes, R. R., Ebert, D., Bonhoeffer, S., 2002. Dose–dependent infection rates of parasites produce the Allee effect in epidemiology. Proceedings
 of the Royal Society of London. Series B: Biological Sciences 269, 271–279.
- Renwrantz, L., 1986. Lectins in molluscs and arthropods: their occurrence, origin and roles in immunity. Symposium of the Zoological Society of
 London 56, 81–93.
- 979 Richardson, L. L., 2004. Black band disease. In: Rosenberg, E., Loya, Y. (Eds.), Coral health and disease. Springer-Verlag, Berlin, pp. 325–336.
- Riisgård, H. U., Larsen, P. S., 1995. Filter-feeding in marine macro-invertebrates: pump characteristics, modelling and energy cost. Biological
 Reviews of the Cambridge Philosophical Society 70, 67–106.
- Riisgård, H. U., Thomassen, S., Jakobsen, H., Weeks, J., Larsen, P. S., 1993. Suspension feeding in marine sponges *Halichondria panicea* and *Haliclona urceolus*: effects of temperature on filtration rate and energy cost of pumping. Marine Ecology Progress Series 96, 177–188.
- Sebens, K., Grace, S., Helmuth, B., Maney Jr, E., Miles, J., 1998. Water flow and prey capture by three scleractinian corals, *Madracis mirabilis*,
 Montastrea cavernosa and *Porites porites*, in a field enclosure. Marine Biology 131, 347–360.
- Sebens, K., Vandersall, K., Savina, L., Graham, K., 1996. Zooplankton capture by two scleractinian corals, *Madracis mirabilis* and *Montastrea cavernosa*, in a field enclosure. Marine Biology 127, 303–317.

- Smith, O. R., 1953. Observations on the rate of decay of soft-shell clams (Mya arenaria). Ecology 34, 640–641.
- Sokolow, S. H., Foley, P., Foley, J. E., Hastings, A., Richardson, L. L., 2009. Editor's choice: disease dynamics in marine metapopulations:
 modelling infectious diseases on coral reefs. Journal of Applied Ecology 46, 621–631.
- Soniat, T. M., Brody, M. S., 1988. Field validation of a habitat suitability index model for the American oyster. Estuaries 11, 87–95.
- Soniat, T. M., Klinck, J. M., Powell, E. N., Hofmann, E. E., 2012. Understanding the success and failure of oyster populations: periodicities of
 perkinsus marinus, and oyster recruitment, mortality, and size. Journal of Shellfish Research 31 (3), 635–646.
- Strathmann, R. R., 1990. Why life histories evolve differently in the sea. American Zoologist 30, 1997–207.
- ⁹⁹⁵ Su, M., Hui, C., 2011. The effect of predation on the prevalence and aggregation of pathogens in prey. Biosystems 105, 300–306.
- Veale, L. O., Hill, A. S., Brand, A. R., 2000. An *in situ* study of predator aggregations on scallop (*Pecten maximus*) dredge discards using a static
 time-lapse camera system. Journal of Experimental Marine Biology and Ecology 255, 111–129.
- Vega Thurber, R. L., Burkepile, D. E., Fuchs, C., Shantz, A. A., McMinds, R., Zaneveld, J. R., 2014. Chronic nutrient enrichment increases
 prevalence and severity of coral disease and bleaching. Global Change Biology 20, 544–554.
- Villalba, A., Reece, K. S., Camino Ordás, M., Casas, S. M., Figueras, A., 2004. Perkinsosis in molluscs: a review. Aquatic Living Resources 17, 411–432.
- Wang, S., Peatman, E., Liu, H., Bushek, D., Ford, S. E., Kucuktas, H., Quilang, J., Li, P., Wallace, R., Wang, Y., et al., 2010. Microarray analysis of gene expression in eastern oyster (*Crassostrea virginica*) reveals a novel combination of antimicrobial and oxidative stress host responses after dermo (*Perkinsus marinus*) challenge. Fish and Shellfish Immunology 29 (6), 921–929.
- 1005 Wang, W., Liu, H., Li, Z., Guo, Z., Yang, Y., 2011. Invasion dynamics of epidemic with the Allee effect. Biosystems 105, 25–33.
- Widdows, J., Brinsley, M., Bowley, N., Barrett, C., 1998. A benthic annular flume forin situmeasurement of suspension feeding/biodeposition rates
 and erosion potential of intertidal cohesive sediments. Estuarine, Coastal and Shelf Science 46, 27–38.
- Widdows, J., Lucas, J. S., Brinsley, M. D., Salkeld, P. N., Staff, F. J., 2002. Investigation of the effects of current velocity on mussel feeding and
 mussel bed stability using an annular flume. Helgoland Marine Research 56, 3–12.
- Wilson–Ormond, E. A., Powell, E. N., Ray, S. M., 1997. Short–term and small–scale variation in food availability to natural oyster populations:
 food, flow and flux. Pubblicazioni della Stazione Zoologica di Napoli I: Marine Ecology 18, 1–34.
- Yakob, L., Mumby, P. J., 2011. Climate change induces demographic resistance to disease in novel coral assemblages. Proceedings of the National
 Academy of Sciences of the United States of America 108, 1967–1969.
- Zvuloni, A., Artzy-Randrup, Y., Stone, L., Kramarsky-Winter, E., Barkan, R., Loya, Y., 2009. Spatio-temporal transmission patterns of black-band
 disease in a coral community. PLoS One 4, e4993.

Parameter	Definition	Unit	Values	Key references
	Disassa transmission rate by filtration or contact with par			Hofmonn at al. (1005):
β_S	ticles.	Particle ⁻¹ time ⁻¹	Eq. 7	Ford et al. (1999)
β_F	Particle removal rate from F corresponding to the infection of S	Particle ⁻¹ time ⁻¹	Eq. 8	Hofmann et al. (1995); Ford et al. (1999)
α	Transmission rate coefficient (see Eqs.7,8)	Particle ⁻¹ time ⁻¹	$2 \cdot 10^{-1}$	Hofmann et al. (1995); Ford et al. (1999); Sokolow et al. (2009)
m_S	Natural mortality rate	time ⁻¹	3.10-4	Powell et al. (2008, 2009); Sokolow et al. (2009)
m_I	Disease mortality rate	time ⁻¹	$8 \cdot 10^{-4}$	Bushek et al. (2012); Choi et al. (1989); Sokolow et al. (2009)
d	Removal rate of dead individuals by scavengers or decay	time ⁻¹	$5 \cdot 10^{-1}$	Hoese (1962); Diamond (2012) Bowell at al. (2008
n _S	Recruitment rate of S	time ⁻¹	$1 \cdot 10^{-2}$	2009); Sokolow et al. (2009)
n_I	Recruitment rate of <i>I</i>	time ⁻¹	$1 \cdot 10^{-2}$	Powell et al. (2008, 2009); Sokolow et al. (2009)
K	Population carrying capacity	Individuals	300	Powell et al. (2012b); Yakob and Mumby (2011)
b_I	Average P per I	time ⁻¹	1.10^{6}	Choi et al. (1989)
b_{DI}	Average IP per dead animal	time ⁻¹	1.10^{7}	Choi et al. (1994); Pow- ell et al. (1996)
b_{min}	Minimum body burden to become I	time ⁻¹	200	Adapted from Ford et al. (1999), Guo, unpub- lished data (ovsters)
b_{max}	Maximum body burden to be considered I	time ⁻¹	400	400
CS	Release rate of pathogens from S	time ⁻¹	$7 \cdot 10^{-3}$	Bushek et al. (2002) Bushek et al. (2002):
c_I	Release rate of pathogens from I	time ⁻¹	$7 \cdot 10^{-3}$	Sokolow et al. (2002),
c _{DI}	Release rate of pathogens from DI	time ⁻¹	$5 \cdot 10^{-1}$	Hoese (1962); Diamond (2012) Sokolow et al. (2009)
r	Loss rate of waterborne pathogens from the local pool	time ⁻¹	$5 \cdot 10^{-2}$	and this study, experi- mental data
$f_S(active)$	Filtration rate of infective particles by S	Individual ⁻¹ time ⁻¹	$1 \cdot 10^{-1}$	Powell et al. (1992)
$f_I(active)$	Filtration rate of infective particles by I	time ⁻¹	$1 \cdot 10^{-1}$	Powell et al. (1992)
$f_H(active)$	Filtration rate of infective particles by the alternate host	Individual ⁻¹ time ⁻¹	$1 \cdot 10^{-2}$	Ben-Horin, unpublished data (tunicates)
f_S (passive)	Contact rate of susceptible individuals with particles	Area ⁻¹ Time ⁻¹	$1 \cdot 10^{-1}$	Adapted from Sokolow et al. (2009)
$f_I(passive)$	Contact rate of I with particles	Area ⁻¹ Time ⁻¹	$2 \cdot 10^{-1}$	Adapted from Sokolow et al. (2009)
$f_H(passive)$) Contact rate of non-focal hosts with particles	Area ⁻¹ Time ⁻¹	$1 \cdot 10^{-1}$	Adapted from Sokolow et al. (2009)
A_S	Exposed particle collection area per S or I	Area Individual ⁻¹	$1 \cdot 10^{-4}$	Adapted from Sebens et al. (1996, 1998)
A_H	Exposed particle collection area per alternate host	Area Individual ⁻¹	$1 \cdot 10^{-4}$	Adapted from Sebens et al. (1996, 1998)
а	Reduction rate of pathogens inside hosts by diapedesis, phagocytosis, apoptosis, etc.	time ⁻¹	$1 \cdot 10^{-1}$	Ben-Horin, unpublished data data (oysters)
sl	Local volume factor to normalize concentration of pathogens; it is the reciprocal to the height of the local volume V_l	vol ⁻¹	10	Wilson–Ormond et al. (1997)
sr	Remote volume factor to normalize concentration of pathogens; it is the reciprocal to the height of the remote volume V_U	vol^{-1}	1	Model calibration
γ	Exchange rate of free-living pathogens between pools	time ⁻¹	$5 \cdot 10^{-1}$	Model calibration
σ	Free-living pathogen lost rate from the remote pool	time ⁻¹	$5 \cdot 10^{-2}$	Ben-Horin, unpublished data (oysters)

Table 2: Parameters of the model. These parameters values are considered plausible for active suspension feeders such as bivalves (e.g. oysters, clams) and passive suspension feeders such as corals, and were adapted from literature data and experimental results obtained for Eastern oyster by one of the authors (Ben-Horin, unpublished data), and from model calibration. Note that the model has an implicit surface area or volume for the parameters. The units used in the simulations are days for time, m^2 for area, and m^3 for volume.